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# GEOCHEMISTRY AND MINERALOGICAL ASSOCIATION OF HEAVY METAL CONTAMINANTS IN FINE GRAINED SEDIMENT, WELLAND RIVER, SOUTHERN ONTARIO.

by

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### Abstract

This investigation of geochemistry and mineralogy of heavy metals in fine grained (<63µm) sediment of the Welland River was undertaken to: 1) describe metal dispersion patterns relative to a source, identify minerals forming and existing at the outfall region and relate sediment particle size to chemistry; 2) to delineate sample handling, preparation and evaluate, modify and develop analytical methods for heavy metal analysis of complex environmental samples.

A joint project between Brock University and Geoscience Laboratories was initiated to test a contaminated site of the Welland River at the base of Atlas Speciality Steels Co. Methods were developed and utilized for particle size separation and two acid extraction techniques: 1) Partial extraction; 2) Total extraction.

The mineralogical assessment identified calcite, dolomite, quartz and clays. These minerals are typical of the carbonate-shale rock basement of the Niagara Peninsula. Minerals such as, mullite and ferrocolumbite were found at the outfall region. These are not typical of the local geology and are generally associated with industrial pollutants.

Partial and total extraction techniques were used to characterize the sediments based on chemical distribution, elemental behaviour and analytical differences. The majority of elements were lower in concentration in the partial extraction technique; suggesting these elements are bound in an acid extractable phase (exchangeable, organic and carbonate phases). The total extraction technique yielded higher elemental concentrations taking difficult oxides and silicates into solution.

Geochemical analyses of grain size separates revealed that heavy metal (Co, Ni, V, Mn, Fe, Ba) concentrations did not increase with decreasing grain size. This is a function of the anthropogenic mill scale input into the river. The background elements (Sc, Y, Sr, Mg, Al and Ti) showed an increase in concentration to the finest grain size suggesting that it is directly related to the local mineralogy and geology.

Dispersion patterns of metals fall into two distinct categories: 1) the heavy metals (Co, Cu, Ni, Zn, V and Cr), and 2) the background elements (Be, Sc, Y, Sr, Al and Ti). The heavy metals show a marked increase in the outfall region, while the background elements show a significant decrease at the outfall. This pattern is attributed to a "dilution effect" of the natural sediments by the anthropogenic mill scale sediments. Multivariant statistical analysis and correlation coefficient matrix results clearly support these results and conclusions.

These results indicate the outfall region of the Welland River is highly contaminated with to heavy metals from the industrialized area of Welland. A short distance downstream, the metal concentrations return to baseline geochemical levels. It appears, contaminants rapidly come out of suspension and are deposited in close proximity to the source. Therefore, it is likely that dredging the sediment from the river may cause resuspension of contaminated sediments, but may not distribute the sediment as far as initially anticipated.



### Acknowledgements

I would like to thank my supervisors Dr. Uwe Brand and Ian Brindle for their direction and constructive comments throughout this study. I would also like to express my sincere thanks Dr. Jean Richardson for her support, guidance, encouragement and patience during the course of this study.

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# Chapter 1: Introduction

# 1.1: Background

Sediments are important in the hydrocycle due to their ability to carry metals. As such, they reflect the current and historical geochemistry of a region (Salomons and Forstner, 1984). Many studies emphasize the correlation between the geochemistry of the finer grained materials and the trace metal contamination in lake and river sediments (Moore et al., 1989; Salomons and Forstner, 1984; Forstner, 1982; Thorne and Nickless, 1981; Ackemann, 1980; Jenne et al., 1980; Filipek and Owen, 1977; Wilber and Hunter, 1979; Gibbs, 1977; Helmke et al., 1977; Whitney, 1975; and Oliver, 1973). Others investigate the distribution of sediment-associated contaminants in waterways (Mudroch, 1984, 1985, Young et al., 1987, Mudroch and Stone, 1989, Gatz et al., 1989, Crecelius et al., 1991, and Bodur and Ergin, 1994). Higher metal concentrations occur in fine grain size fraction sediments (<63 µm) due to their large surface area/volume ratio. Often, this metal occurrence is attributed to adsorption, coating, coprecipitation or complexing.

Grain size separation is beneficial in local environmental studies because only a few samples are essential. Samples consisting of fractions greater than 173 µm material contain a high percentage of large grains and are generally not high in metals (Salomons and Forstner, 1984). Typically, the <63 µm fraction is recommended for contamination studies because trace metals are mainly associated with silt and clay size particles. This fraction is nearly equivalent to the size of material carried in suspension, which is known to transport a high percentage of contamination (Forstner and Salomons, 1980). Metal concentration is not altered by sieving and numerous metal studies are performed on the <63 µm fraction,

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allowing better comparison of results (Forstner and Salomons, 1980) The ability to analytically process samples of a highly complex nature is important now and will become critical in the future, as a better understanding of chemical species in environmental studies become known, the need for lower detection limits will be essential.

To obtain a comparative basis for environmental studies, the background concentrations of trace metals should be established in similar sample media (Forstner and Wittmann, 1979). There are four concepts that can be employed to achieve this: 1) average shale composition as a global standard value; 2) fossil aquatic sediments from defined environments as a standard, (taking into account natural allochthonous and autochthonous factors and mechanisms and regional influences); 3) recent deposits in relatively unpolluted areas; and 4) short, dated sediment cores, which provide a historical record of events occurring in the watershed of a particular river. The background concentration of major and trace elements in the suspended load is obviously affected by the composition of soils and bedrock geology of the drainage basin and varies over the local area (Mudroch and Duncan, 1986).

Often, environmental studies will involve chemical extractions of specific grain sizes of sediment. Metals in sediments are concentrated by physical (grain size, surface area) or chemical (cation exchange capacity, mineralogy, concentration of geochemical substrates) factors (Horowitz et al., 1989). By employing both of these concepts, the geochemistry of sediments can provide insights into processes that affected the sediment-trace element distributions. Changes in ambient conditions of the environment could cause mobilization of metals. Subsequently, the metals may be available to the organisms in the environment.

Often trace metals are present in the sediment as surface "coatings". Metals in such coatings are the easily extracted, and therefore, readily available for solution testing (Simon et al., 1992).

The objectives of this thesis are to evaluate, modify and develop analytical methods for heavy metal analysis of complex environmental samples, relate sediment particle size to chemistry, and to describe metal dispersion patterns relative to a source region in the Welland River. Finally, from this information, minerals forming and existing in the Welland River will be assessed and metal presence and dispersion patterns explained.

# 1.2: General Geology and Study Area

The Welland River is located in the Niagara Peninsula. Starting along a sandy moraine near Ancaster, it meanders across the Haldiman Clay Plain where it used to drain into the Niagara River at Chippawa, south of Niagara Falls, Ontario (Fig. 1). The Welland River actually does not flow into the Niagara River any longer. The last four miles of the river serves as an intake for the Chippawa-Queenston power canal. In addition, further downstream, the river is forced through two siphons under the Welland Canal in the City of Welland (Chapman and Putnam, 1984). The river has a high suspended sediment load (Brindle, personal communication). The water has a relatively high pH (average 7.8, Appendix 3), which is a reflection of the carbonate rock environment which consists of dolostones of the Lockport Formation, capping the local bedrock.

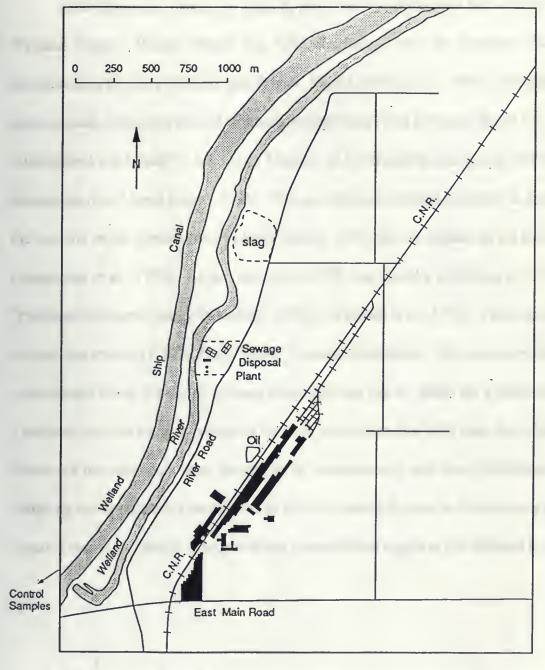


Fig.1 Sample area of the Welland River located in Welland, Niagara Peninsula.



The overburden on the Paleozoic rocks ranges from 50-150 feet thick, increasing in a southerly direction (Chapman and Putnam, 1984).

Atlas Speciality Steels Co. (Fig. 3) and other industrial sites are located on the Waste effluent has been discharged into the Welland River for Welland River. approximately 60 years (Thomas and Palmer, 1989; Dickman et al. 1990). Recognition of contamination of the river has led to a commitment from Atlas speciality Steels Co., Acres International (on behalf of Atlas) and Ministry of Environment and Energy (MOEE) to cleanup the river (Acres Report, 1991). This commitment resulted in studies to determine the severity of the contamination (Acres Report, 1991) and the impact on the local biota (Jaagumagi et al., 1995). At this time, the MOEE had issued a preliminary draft of the "Provincial Sediment Quality Guidelines" (PSQG) (Persaud et al., 1993). These new limits replace the ministry's 1976 Open Water Disposal Guidelines. This document contains contaminant levels for which cleanup considerations can be based for a particular site. Therefore, for this study, a portion of the river was selected in 1991 near Atlas Specialty Steels, as the sediments were thought to be contaminated with heavy metals and oils. Sampling was completed in the summer of 1991 (Appendix 5), prior to the dredging project. Figure 2 shows the sample locations in the industrialized region of the Welland River.

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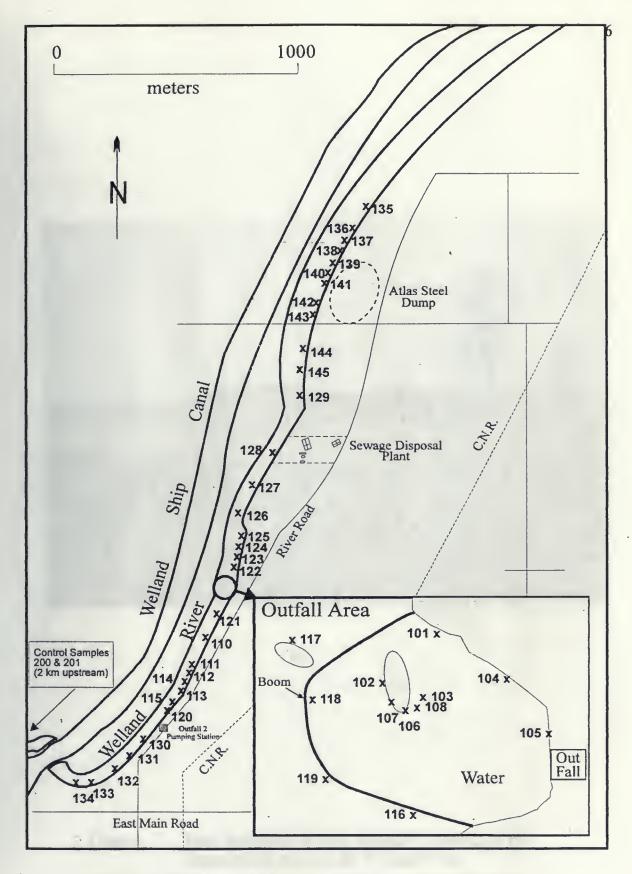


Figure 2: Field map of sample localities on the Welland River.



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Figure 3. Field photograph of Atlas Speciality Steels Co. in the industrialized region of the Welland River.



### 1.3: Mineralogical Method Description

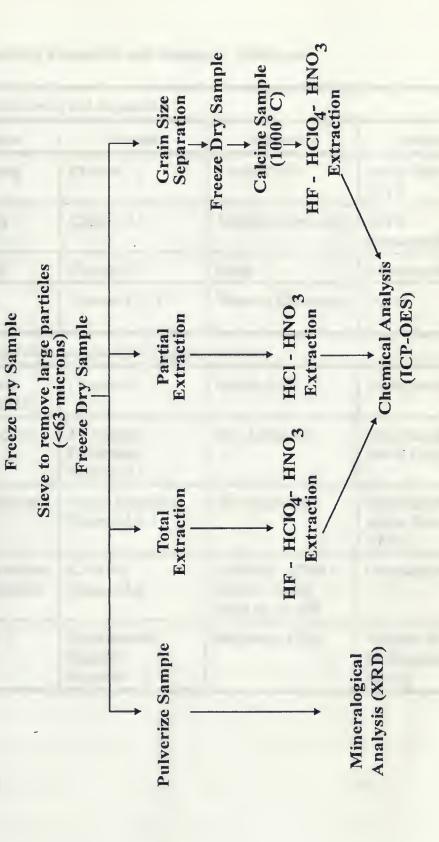
Mineralogical analyses were performed on the samples to identify unique mineral existing or forming in the Welland River. Traditional X-ray diffraction (XRD) analyses were carried out at Geoscience Laboratories on the samples. Details of the technique and instrumentation are available in Chapter 2.6 - Mineralogical Analysis and are available at Geoscience Laboratories (Lab Manual ref. MP 149).

### 1.4: Geochemical Method Descriptions

In the analysis of trace metals for environmental geochemical studies, a partial extraction or HCl-HNO<sub>3</sub> leach may be used to identify metals available to the surrounding environments such as plants, animals, and in surface or ground waters. In contrast, a total digestion technique may be used to determine all metals in the sample; those available from the natural/geological environment and those introduced as contamination.

Investigation into methodology, equipment availability and applicability have resulted in the delineation of a scheme for sampling and for analytical procedures (Fig. 4). These procedures were extensively tested and proved optimal for the study. More detail on the procedures and techniques is available in Chapter 2.4 - 2.5. In this study, a number of analytical preparation and instrumentation techniques were used at several institutes. These are tabulated in Table 1.

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Sample Collection

Freeze Sample

Figure 4. Flowchart delineating the analytical procedures required for sample pretreatment, preparation and chemical analysis.

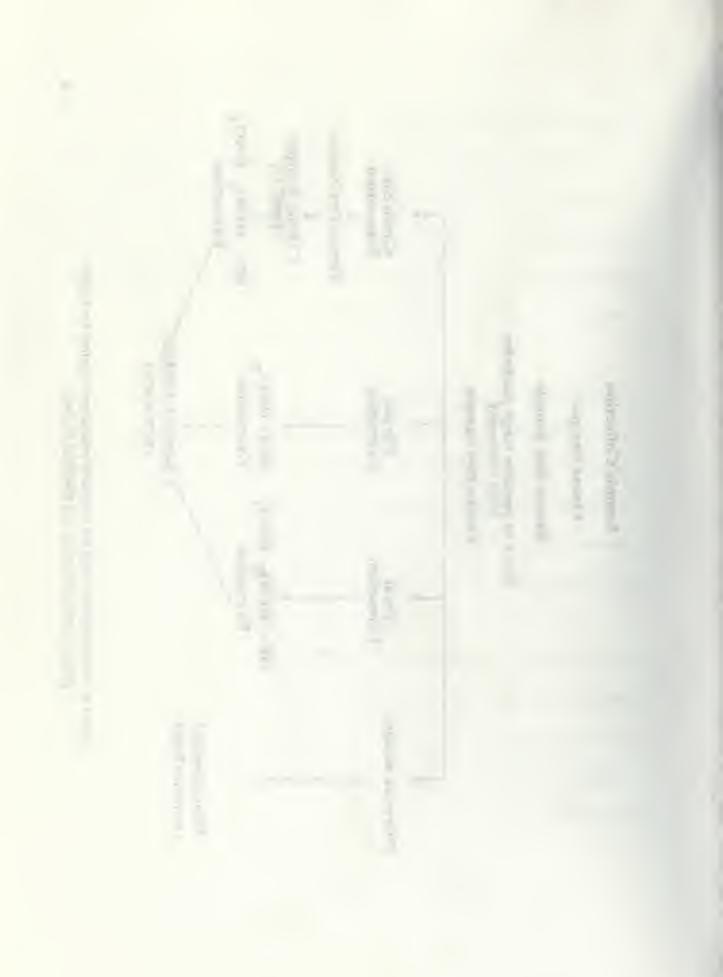


Table 1: Analytical Preparation and Instrument Techniques

Sample Pretreatmen	t and Separation			
Technique	Chapter	Instrument	Institution	
Freeze Drying	Chapter 2.2	Freeze Dryer	Brock University CCIW	
Wet Sieving	Chapter 2.2	Stainless steel sieve	CCIW Geoscience Labs	
Dry Sieving	Chapter 2.2	Rotap	Geoscience Labs	
Grain Size Separation	Chapter 2.2, 2.3	Warman Cyclosizer	CCIW	
Sample Preparation and Analytical Instrumentation				
Mineralogical Analysis (XRD)	Chapter 2.4	Philips XRD	Geoscience Labs	
Total Extraction	T2 Sample Preparation Chapter 2.5.1	Not Applicable	Geoscience Labs Brock University	
Partial Extraction	Partial Extraction Chapter 2.5.2	Not Applicable	Geoscience Labs advise from XRAL	
Chemical Analysis & Instrumentation	ICP-OES Chapter 2.6	ICP-OES - JY48+, Thermo Jarrell Ashe ICAP 61E	Geoscience Labs	
Microwave Digestion	Experimental Digestion Appendix 2	Milestone 1200	Ministry of Environment & Energy	



### Chapter 2: Sample Pretreatment and Analytical Methods

### 2.1: Sample Collection

A total of 50 bottom sediment samples, including two control samples, were collected from the Welland River (Fig. 1). Grab samples were obtained from the outfall area (Fig. 5), upstream (Fig. 6), and downstream (Fig. 7) from the Atlas-Mansfield discharge point source (Fig. 2). Each sample was located on an airphoto and given a sample label. The samples were collected with the use of a plastic implement to ensure that metal contaminants were not added in the sampling procedure. Wet samples were placed in plastic bags without draining off the excess water. This procedure maintained sample integrity as fine grained material in suspension was not lost. The samples were brought to the laboratory and stored in a freezer until initial pretreatment.

## 2.2: Sample Pretreatment

The samples were taken from the deep freezer, freeze-dried, and stored. A selected set of samples were wet-sieved through a 63µm stainless-steel sieve to separate the coarse material from the finer material of interest. The coarse fraction (>63µm) was stored and the fine-grained material was freeze-dried with the exception of the samples that would be further separated into six grain sizes (<11µm, 11-15µm, 15-23µm, 23-33µm, 33-44µm, 44-63µm) following the procedure used by the Canada Centre for Inland Waters (CCIW), then freeze-dried and stored in vials. Another set of samples were dry sieved using a stainless steel sieve and Rotap to separate samples mechanically to less than 63µm grain size. This set was completed in this manner because of the speed of the method.

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Figure 5. Field photograph of the outfall region (Atlas-Mansfield storm sewer) of the Welland River.





Figure 7. Field photograph of the Welland River downstream from the Atlas-Mansfield discharge point source.



Figure 6. Field photograph of the Welland River upstream from the Atlas-Mansfield discharge point source.



### 2.3: Grain size Separation Methods

Several experiments were performed to assess the method most feasible for grain size separation below 63 µm. The method used by Geoscience Laboratories was more appropriate for larger grain size than the finer grain sized (below 63 µm) material of interest in this study. The use of sodium hexametaphosphate could interfere with trace metal analysis. Dry sieving the samples failed because the mesh was too fine and the sediment did not pass through the sieve easily. Finally, a Warman Cyclosizer was used to separate particles by centrifuging in the sub-sieve range. A suite of 18 samples were randomly chosen to encompass a distribution along the river. Each sample was wet sieved, the <63 µm fraction was introduced to the Cyclosizer. The Cyclosizer separated the sample into five grain size ranges, the sixth sample range (<11 µm) was obtained by allowing the waste water and sediment from the instrument to settle in a large container for 48 hours. The limiting particle separation size at standard conditions and the conditions at which the Cyclosizer operates are summarized in Table 2.

Once the sample was separated in each cyclone it was decanted into a clean beaker and allowed to settle. After settling, the grain size split was transferred to a plastic vial for freeze-drying.

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Table 2.: Cyclosizer operating conditions and limiting particle separation size.

Cyclone No.	Particle Limit	Error (μm)
1	44 microns	±2
2	33 microns	±1
3	23 microns	±1
4	15 microns	±1
5	11 microns	±1
Conditions		3
Water Flowrate	11.6 L/min	
Water Temperature	20°C	
Particle Density	2.65 g/cm <sup>3</sup>	
Time of elutriation	infinite	

# 2.4: Mineralogical Analysis

# 2.4.1: Sample Preparation

Sediments were pulverized using an agate mortar and pestle until a fine powder was obtained. The sample was poured into a special aluminum well mount. A glass slide was held in front of the ring and tapped until the powder was packed down. This method avoided preferred orientation of minerals. This preparation is a non-destructive method, so all samples were retained after XRD analysis.

### 2.4.2: Instrumentation

Mineralogical analysis was performed by Phillips X-ray Diffraction (XRD) system. The powder mount was placed into the diffractometer and held by a spring clip for even illumination by the X-ray beam. The diffractometer analyzes ranges from 5° to 65°θ. Once the analysis was complete, the diffraction pattern was printed and compared to the JCPDS search database to match their peaks to a mineral. Details of the techniques and procedures are outlined in OGS 1990.

### 2.5: Digestion Techniques

### 2.5.1: Total Extraction Technique

An initial attempt at a total extraction procedure EA18 (T2) (OGS, 1990) method resulted in an incomplete digestion with a large amount of residue. Several modifications of this procedure were tested to digest the samples as completely as possible and yield the most accurate and precise data attainable. The acids were introduced to the samples at varying times, but this yielded no significant difference in the completeness of the digestion. Since oil was observed in samples during sample collection, it was possible that the residue was an organic material. Accordingly, H<sub>2</sub>O<sub>2</sub>(hydrogen peroxide) was added to the digestion. However, residues were still present. Finally, the sample was calcined at 1000°C prior to acid digestion to combust oils and volatiles (CO<sub>2</sub>, S, H<sub>2</sub>O, As, Hg, Se). This resulted in a much cleaner and complete digestion. This modification was tested using Standard Reference Materials to assess the effect on the analytes to be determined.

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A total of 52 river samples were analyzed using a total extraction technique. A weight of 0.5g of sample was placed in a PTFE beaker. Normally a mixed acid attack of HNO<sub>3</sub>- HClO<sub>4</sub>-HF (1:2:8) yielded a total decomposition of samples to produce a analytical solution. An amount of 15 mL of acid was added to the beaker and the contents gently swirled to ensure wetting of the entire sample. The beakers were placed on a hot plate at 150°C. Digestion took approximately 20 hours and eventually the sample was evaporated to dryness. Once dry, the beakers were gently tapped to break the residue and cause any droplets of acid around the top portion of the beaker to fall to the bottom. The beakers were returned to the hot plate until fuming ceased. This step was repeated until the sample was completely dry. Subsequently, 2 mL of HNO3 was added to the dry residue and heated for 1 minute on the hot plate. The sample was removed, 0.5 mL of HCl added and the beaker returned to the hot plate once the reaction had subsided for approximately 1 minute. The residue should appear to be dissolved in the acid solution. At this point, 15 mL of deionized water was added to the beaker and heated for 15 minutes. Record of any incompletely dissolved residue was indicated on the worksheet. Once the solutions cooled, they were transferred to 50 mL graduated cylinders and brought up to volume with deionized water. The cylinder was covered with parafilm and mixed. The solution was transferred to a bottle or test tube to await ICP-OES analysis. The dilution factor for ICP-OES analysis was 100. Occasionally, if a sample yielded a large amount of residue, an XRD scan was performed to identify the mineral phases not completely digested in the acid attack.

Microwave digestion was also tried hoping to yield a total digestion. HNO<sub>3</sub> and HCl were used in the attack. However, HF could not be introduced to the bombs because of MOEE safety requirements. The digestion that resulted was not a total extraction and was very time consuming (12 samples per day), so the procedure was abandoned. See Appendix 2.

### 2.5.2: Partial Extraction Technique

Several experiments were performed to assess the most precise and accurate partial extraction technique to be used. Initially, HNO<sub>3</sub> was added and HCl later. This resulted in erratic analytical results. Varying temperature and digestion times yielded limited differences. However, adding HCl first and then HNO<sub>3</sub> later and increasing the temperature to 90°C resulted in consistently precise and accurate (within ±20%) data (Obdebeck, XRAL, personal communication; Scriver et al., 1995).

A total of 52 river sediments were analyzed using a partial extraction technique. A weight of 0.25g of river sediment of less than 63µm was placed into a glass test tube with 2 mL of HCl and placed in a water bath at 80-90°C for 30 minutes. Subsequently, 1 mL of HNO<sub>3</sub> was added and the sample returned to the water bath for 2.5 hours. Each tube was shaken using a Vortex shaker every 20 minutes. The test tubes were cooled and solutions transferred to 25 mL volumetric flasks and brought up to volume with deionized water.

### 2.6: Instrumental Analysis of Samples by Partial and Total Extraction

### 2.6.1: Instrumentation

ICP-OES (Inductively Coupled Plasma-Optical Emission Spectrometry) analysis is an effective way of determining most elements on the Periodic Table. Most analytes exhibit a linear response over five orders of magnitude, with limits of detection in the low ppm range (Thompson and Walsh, 1989). Figure 8 shows a schematic diagram of a conventional simultaneous ICP-OES system. A sample is introduced to the plasma as an aqueous aerosol via a nebulizer, passes through a spray chamber to remove large droplets, then injected into the plasma (Greenfield and Montaser, 1992). The atoms or ions are energized, and when reverting to a lower energy state emit a photon of energy. This photon is then converted to an electrical signal by the photomultiplier in the spectrometer. The intensity of the electrical signal is compared to that of previously measured known solutions and consequently, a concentration can be calculated (Thompson and Walsh, 1989).

# 2.6.2: Geochemical analysis by ICP-OES

Analysis was done by ICP-OES because of its analytical range (low ppm to mid percent levels). This analytical range was found to be the range required for these samples and for the analytes in question. ICP-OES also has the capability of analyzing several elements at a time. This was important due to the small sample size obtained from the grain size splits. Often there was only enough sample for one preparation. Therefore, it was crucial to be able to analyze for all analytes simultaneously.

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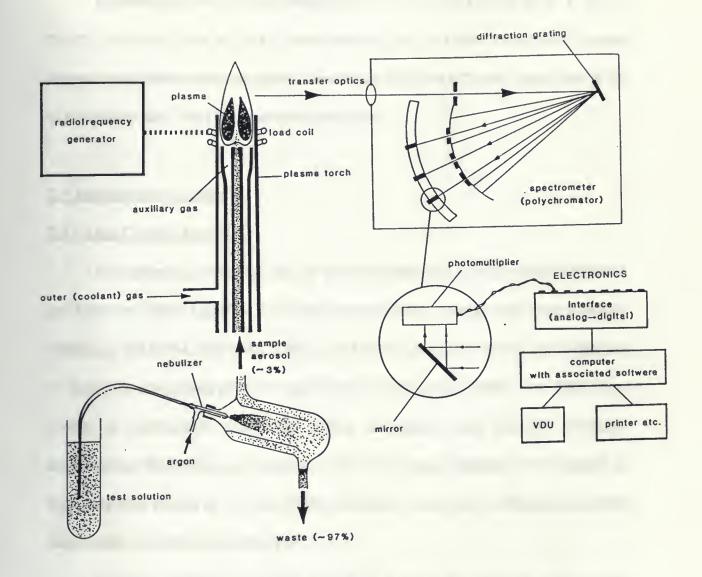
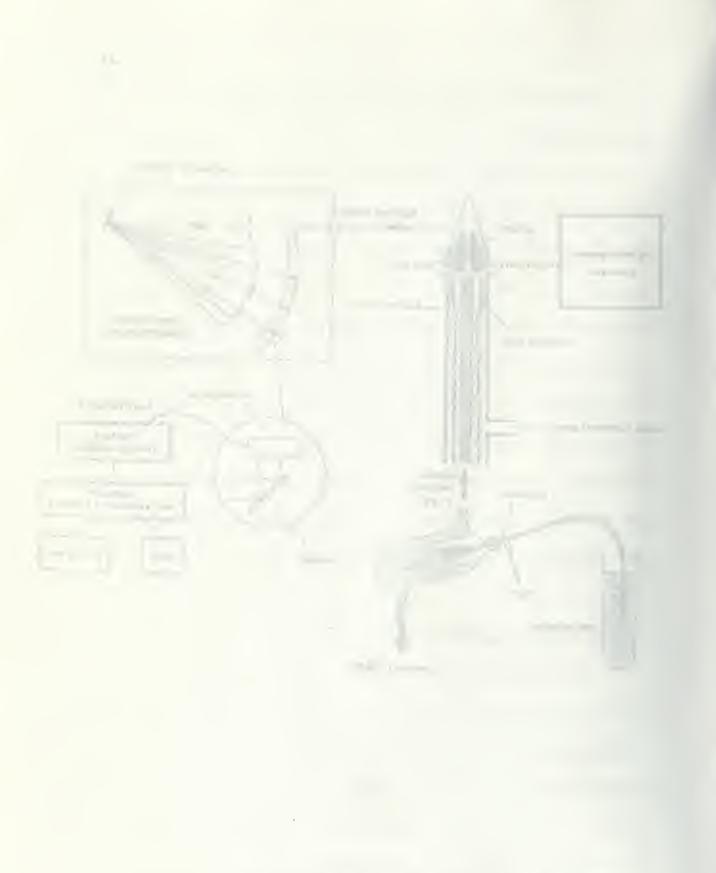


Figure 8. Schematic diagram of a conventional simultaneous ICP-OES system (Thompson and Walsh, 1989)



Preliminary results and experimental data for Be, Co, Cu, Mo, Ni, Sc, V, Y, Sr, W, Zn, Ti, Al, Fe, Mg, Mn, Ba and Cr were obtained using a JY48+ Inductively Coupled Plasma-Optical Emission Spectrometer (Toronto, pre 1992) and a Thermo Jarrell Ashe ICAP 61E (Sudbury post 1992) at Geoscience Laboratories.

### 2.7: Precision and Accuracy

### 2.7.1: Data Quality Objectives

Evaluation of geochemical data are well established and are outlined in Richardson and Morrison (1995), Lightfoot (1993) and Johnson (1993). Standard reference materials (SRM) (i.e. NIST2704, NIST1646, STSD-2 and LKSD-3) were used in each set of analyses to determine appropriate precision and accuracy (Govindaraju, 1989). The SRM were chosen to approximate closely the matrix, composition and expected elemental concentrations (Richardson and Morrison, 1995) of the stream sediments. The elements in this study were chosen on the basis of their elemental background representation, health implications and environmental impact.

Samples, standards and blanks were used to assess the precision, accuracy and laboratory technique contamination levels of all the data. A recoverability of 80% of the certified value was achieved in the partial extraction data and total extraction data. An assessment of the quality control samples was to determine technique contamination, precision and accuracy in all batches.

### 2.7.2: Analytical Considerations

Concentrations of some elements in the SRM used were less than Method Detection Limit (MDL) of the method (ICP-OES). e.g. Mo - MDL < 6ppm, W - MDL < 35ppm, where the only alternative to this problem was to use a more sensitive method. Before acceptably good accuracy is possible, one must have acceptable precision. If data are between Method detection limit and Limit of Quantification (LOQ), [where the LOQ = 3.3 x MDL], the precision will be poor (ie. Tungsten has a method detection limit of 35ppm, therefore, the precision is worse than ±100% because it cannot be measured well). The only alternative to improve this is to use a more sensitive method or increase the sample size. (See Fig. 9).

Some of the duplicates do not agree well, probably due to sample inhomogeneity (Table 3). Each sample was sieved to <63 µm, then, in turn, divided into grain size splits A through F. These splits are of the bulk samples (<63 µm, primary focus of the study) and are not representative of the sampling environment. Thus, they do not yield the same volume subsample uniformly, hence the potential for these splits to be inhomogeneous is high. Small sample size is known to contribute to this effect.

The potential of spattering of extremely high concentration samples into low concentration samples resulted in some ambiguity in elements with low sample concentrations. Overall, the anthropogenically contaminated sediments have elemental concentrations 1 to 4 orders of magnitude higher than those of the background samples and the SRM.

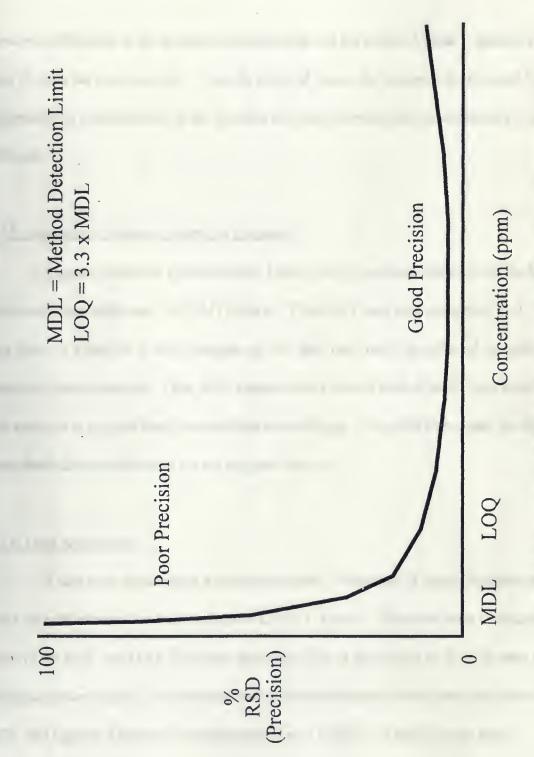


Figure 9. Precision of data as concentration approaches the method detection limit.



However, differences in the measured concentration and the certified value is generally less than 20 ppm for most analytes. Thus, in terms of fitness for purpose, the elevated levels differences in concentration in the samples are negligible and can be interpreted without difficulty.

#### 2.7.3: Evaluation Criteria - Fitness for Purpose

A Highest Analytical Contamination Level (HACL) was established from the blank values and "ppm difference" of SRM (Table 4). This HACL was then multiplied by 2. This was done to establish a safety margin on the data that might be affected by potential laboratory contamination. Thus, each element had a cut-off limit at which low level data was accepted or rejected based on analytical methodology. All results higher than the HACL were deemed acceptable data for the purpose the study.

#### 2.7.4: Data Assessment

All data were separated on a per-element basis. Evaluation of blanks, duplicates and SRM for each element are summarized in Tables 3, 4 and 5. The data were considered in terms of the MDL and LOQ. Only data above the MDL of the method (ICP-OES) were used. Only background levels (ie. concentrations in samples obtained from upstream) above the MDL and Highest Analytical Contamination Level (HACL) (Table 7) were used.

Table 3: Concentration of reagent blanks and concentration difference and precision calculation of duplicate analyses.

Sample #	Be	Sc	Y	5r	Co	Cu	NI	Cr	v	Zn	Al	Ba	Fe	Mg	Mn	Ti
Blank-1	0	0	0	0	0	11	3	17	1	4	38	4	79	6	1	4
Blank-2	0	0	0	0	0	1	t	2	1	6	41	1	10	0	0	3
Blank-3	0	0	0	0	0	1	1	5	1	1	39	L	25	0	0	4
Blank-4	0	0	0	0	0	0	0	2	1	1	37	1	23	0	0	4
Blank-5	0	0	0	0	0	1	1	1	ı	1	9	0	9	0	0	3
Average	0	0	0	0	0	6	2	9	1	. 2	23	2	44	3	1	3
Sample #	Be	Sc	Y	Sr	Co	Cu	NI	Cr	v	Zn	Al	Ba	Fe	Mg	Mn	TI
108-A	0.06	1.31	2.58	13.62	153.5	776.8	6254	12630	112.6	138.9	3577	91.63	21060	2288	6241	256.4
108-A-DUP	0.06	1.18	2.44	14.05	156.6	854.3	6339	13650	119.1	137.5	3629	89.09	20780	2475	6298	249.0
108A ppm diff	0.00	-0.13	-0.14	0.43	3.10	77.5	85.0	1020	6.50	-1.40	52.0	-2.54	-280	187	57.0	-7.40
108A-percent diff	-2.45	10.14	5.35	-3.16	-2.02	-9.98	-1.36	-8.08	-5.77	1.01	-1.45	2.77	1.33	-8.17	-0.91	2.89
124-A	0.41	1.64	3.20	24.27	120.4	673.7	4780	15090	153.3	138.2	6927	297.4	223400	2656	5097	443.1
124-A-DUP	0.45	1.72	3.38	25.74	125.3	710.4	4874	15640	160.9	139.4	6984	303.6	225600	2675	5037	474.
124A ppm diff	0.04	0.08	0.18	1.47	4.90	36.7	94.0	550	7.60	1.20	57.0	6.20	2200	19.0	-60.0	31.6
124A-percent diff	-8.56	-5.00	-5.46	-6.06	-4.07	-5.45	-1.97	-3.64	4.96	-0.87	-0.82	-2.08	-0.98	-0.72	1.18	-7.13
104-C	1.00	6.74	18.62	210.5	10.78	282.2	69.46	130.9	55.62	172.3	53710	427.2	23760	14480	686.8	3803
104-C-DUP	0.92	6.57	18.26	205.1	9.97	277.7	75.24	138.5	50.39	92.07	52110	434.2	23760	14170	669.7	3841
104C ppm diff	0.08	0.17	0.36	5.40	0.81	4.50	-5.78	-7.60	5.23	80.23	1600	-7.00	520	310	17.10	-38.0
104C-percent diff	8.40	2.55	1.93	2.57	7.50	1.59	-8.32	-5.81	9.40	46.56	2.98	-1.64	2.19	2.14	2.49	-1.00
115-C	2.12	12.91	18.72	176.3	21.51	510.3	176.4	476.9	125.4	268.2	89830	599.3	50880	18680	556.0	5235
115-C-DUP	2.31	13.81	19 77	178.4	22.7	537.5	179.0	442.6	130.0	273.6	89330	588.1	51450	18350	554.7	5263
115C ppm diff	0.19	0.90	1.05	2.10	1.19	27.20	2.60	-34.30	4.60	5.40	-500	-11.2	570	-330	-1.30	28.0
115C-percent diff	-9.05	-6.97	-5.61	-1.19	-5.53	-5.33	-1.47	7.19	-3.67	-2.01	0.56	1.87	-1.12	1.77	0.23	-0.53
125-C	1.65	5.20	8.75	127.8	155.6	1180	7556	25540	275.2	587.8	32320	2499	213900	11650	4157	1645
125-C-DUP	1.60	5.27	8.97	129.7	154.1	1213	7371	25210	279.2	585.1	32620	2565	243000	11520	4331	1696
125C ppm diff	-0.06	0.07	0.22	1.90	-1.50	33.0	-185	-330	4.00	-2.7	300	66.0	29100	-130	174	51.0
125C-percent diff	3.39	-1.42	-2.51	-1.49	0.96	-2.80	2.45	1.29	-1.45	0.46	-0.93	-2.64	-13.60	1.12	-4.19	-3.10
																40.00
127-C	2.01	11.41	16.12	158.6	27.31	543.8	464.4	1757	129.6	387.1	76070	520.2	70880	17670	886.9	4282
127-C-DUP	2.10	11.68	16.39	163.4	27.74	515.9	472.6	1979	132.5	394.4	78410	569.8	72910	18200	922.6	4271
127C ppm diff 127C percent diff	0.09 -4.44	0.27 -2.37	0.27 -1.67	4.80 -3.03	0.43 -1.57	-27.90 5.13	8.20 -1.77	222.0 -12.64	2.90 -2.24	7.30 -1.89	2340 -3.08	49.6 -9.53	2030 -2.86	530 -3.00	35.7 -4.03	-11.0
							2000				0.1.4					
200-C	1.83	11.03	20.53	172.7	15.25	379.6	36.83	60.63	89.34	124.3	77060	528.5	32500	10530	279.6	5027
200-C-DUP	1.91	10.94	19.67	169.8	15.78	382.5	39.35	65.11	90.98	125.6	74640	529.2	32260	10260	276.9	4618
200C ppm diff	0.09	-0.09	-0.86	-2.90	0.53	2.90	2.52	4.48	1.64	1.30	-2420	0.70	-240	-270	-2.70	-409
200C percent diff	-4.66	0.82	4.19	1.68	-3.48	-0.76	-6.84	-7.39	-1.84	-1.05	3.14	-0.13	0.74	2.56	0.97	8.14
101-D	1.92	7.50	17.53	239.9	242.0	2933	10140	22300	361.7	2029	52320	1354	35570	28870	9062	2581
101-D-DUP	1.42	5.58	12.99	171.9	178.8	2027	7537	16310	268.5	1490	37780	1000	27900	21620	6558	1955
101D ppm diff	-0.50	-1.92	-4.54	-68.0	-63.2	-906	-2603	-5990	-93.2	-539	-14540	-354	-7670	-7250	-2504	-626
101D-percent diff	26.26	25.60	25.90	28.35	26.12	30.89	25.67	26.86	25.77	26.56	27.79	26.14	21.56	25.11	27.63	24.25
113-D	2.46	14.02	20.10	176.1	36.83	962.2	558.0	1936	171.2	, 430.8	94820	617.8	74470	21430	677.2	5622
113-D-DUP	2.24	12.50	18.01	150.3	30.97	792.0	473.0	1572	147.1	361.0	77650	507.3	62610	17630	563.6	4556
113D-ppm diff	-0.23	-1.52	-2.09	-25.8	-5.86	-170	-85.0	-364	-24.1	-69.8	-17170	-111	-11860	-3800	-114	-106
113D-percent diff	9.18	10.84	10.40	14.65	15.91	17.69	15.23	18.80	14.08	16.20	18.11	17.89	15.93	17.73	16.77	18.90
123-D	1.82	5.99	10.97	165.5	175.1	1301	9330	28390	288.9	634.5	35810	1740	207200	15940	4145	1850
123-D-DUP	1.92	6.18	11.24	170.5	178.0	1356	9496	28920	295.4	644.3	37140	1240	206700	16420	4206	1912
123D ppm diff	0.10	0.10	0.27	5.00	2.90	55.0	166	530	6.50	9.80	1330	-500	-500	480	61.0	62.0
123D-percent diff	-5.22	-3.29	-2.46	-3.02	-1.66	-4.23	-1.78	-1.87	-2.25	-1.54	-3.71	28.74	0.24	-3.01	-1.47	-3.35
1	0.22	0		0.02	,,,,,											
130-D	2.24	12.92	18.16	157.3	22.59	1048	181.5	385.7	122.6	380.8	84340	564.6	45100	15350	351.8	5190
130-D-DUP	2.25	13.03	18.25	159.6	23.06	1034	182.0	408.3	123.2	379.4	84280	568.8	45430	15410	356.0	5080
130D ppm diff	0.01	0.11	0.09	2.30	0.47	-14.0	0.50	22.6	0.60	-1.4	-60.0 0.07	4.2 -0.74	330	60.0 -0.39	4.20	-110
130D percent diff	-0.31	-0.85	-0.50	-1.46	-2.08	1.34	-0.28	-5.86	-0.49	0.37	0.07	-0./4	-0.73	-0.39	-1.19	2.12
201-D	2.18	13.73	18.99	157.4	16.20	786.9	48.99	64.38	118.5	173.3	88830	557.0	41080	12950	307.9	6008
201-D-DUP	2.27	14.13	19.75	160.1	16.57	802.5	54.71	96.36	123.2	175.5	87950	558.8	41450	12660	311.2	5884
201D-ppm diff	0.08	0.40	0.76	2.70	0.37	15.60	5.72	31.98	4.70	2.20	-880	1.80	370	-290	3.30	-124
201D-percent diff	-3.80	-2.91	-4.00	-1.72	-2.28	-1.98	-11.68	-49.67	-3.97	-1.27	0.99	-0.32	-0.90	2.24	-1.07	2.06

All values in ppm.



Table 4: Variation in concentration of analyses of Standard Reference Materials.

	6	ζ	>	2	ζ	(	1	Ç	1		1	F				Ė
91-1646 ppm diff	0.4	1.2	3.4	2.6	3  %	1 9	<b>3</b> 1 8	にに	19.6	12.7	A10263	73.9	<u>re</u> 8419	MR 1752	38.4	122
92-1646 ppm diff	0.4	1.4	2.9	0.9	2.9	8.4	10.0	3.6	21.5	20.1	9993	77.9	8489	1472	41.7	325
93-1646 ppm diff	0.3	0.1	2.0	13.0	1.1	6.0	3.7	2.7	10.3	2.5	4923	24	4139	832	1.6	425
94-1646 ppm diff	0.3	0.2	2.6	20.4	9.0	0.2	3.7	1.7	7.5	7	1993	38	3319	372	4.6	675
95-1646 ppm diff	0.3	0.30	2.6	20.8	0.1	1.2	4.2	4.7	5.2	0.5	2733	21.7	3809	902	2.9	664
AVERAGE	0.3	9.0	2.7	12.6	1.3	2.2	9	4	12.8	9.7	5981	47.1	5635	1066	18	456
91-2704 ppm diff	NA	0.7	NA	26.1	3.0	20.8	12.5	12.2	24.1	76.3	11752	82.3	9205	2076	95.2	246
92-2704 ppm diff	NA	0.5	NA	9.5	1.0	9.1	4.9	20.2	0.6	23.2	3232	21.3	3505	226	15.3	254
93-2704 ppm diff	NA	0.4	NA	10.7	0.7	6.9	5.8	0.7	11.9	42.3	5442	28	4225	908	36.5	463
94-2704 ppm diff	NA	6.0	NA	3.8	0.7		5.5	0.1	9.8	19.4	2892	42.3	2845	456	21.9	284
95-2704 ppm diff	NA	0.7	NA	5.3	0.5	5.7	4.0	2.6	% %	28.6	5272	33.8	4145	962	34.7	828
AVERAGE	NA	9.0	NA	11.1	1.2	8.9	6.5	7.2	12.5	38.0	5718	41.5	4785	872	40.7	929
93-lksd-3 ppm diff	0.2	0.7	4.4	31.4	3.8	2.8	13.8	9.0	9.01	00 00	5232	32.6	4984	968	29.0	340
95-lksd-3 ppm diff	0.2	6.0	5.0	21.2	3.9	1.5	13.3	5.1	9.4	8.3	2462	34.8	4584	516	14.0	420
AVERAGE	0.2	8.0	4.7	26.3	3.8	2.1	13.6	2.8	10.0	9.8	3847	33.7	4784	206	21.5	380
93-stsd-2 ppm diff	0.7	1.6	6.7	51.5	2.7	7.2	16.1	1.8	15.1	50.1	0229	8.3	1470	1710	4.0	110
94-stsd-2 ppm diff	0.7	1.6	7.2	36.7	3.2	10.0	14.3	14.1	13.0	30.3	6930	24.0	760	1690	13.0	174
95-stsd-2 ppm diff	0.7	1.7	6.9	42.2	3.2	4.5	13.8	15.1	13.0	20.4	4220	10.8	2100	1220	10.0	295
AVERAGE	0.7	1.6	6.9	43.5	3.0	7.2	14.7	10,3	13.7	33.6	5973	14.4	1443	1540	0.6	193

Ali values in ppm. Certified values from primary certificates. 1646 = NIST 1646 - Estuarine Sediment 2704 = NIST 2704 - Buffalo River Sediment Iksd = CANMET Lake Sediment stsd = CANMET Stream Sediment NA = Data not available 1111 75 bbs IN

Table 5: Trueness of several analyses of Standard Reference Materials.

% Accuracy Calculations	tions															
Sample #	Be	Sc	×	Sr	3	리	Z	히	>	Zn	V	Ba	Fe	Mg	Mn	ijΠ
91-1646 percent diff	28.1	11.0	16.3	1.5	16.9	22.2	28.2	9.3	20.9	9.2	16.4	17.9	25.1	16.0	10.3	3.8
92-1646 percent diff	28.2	12.6	13.9	3.5	27.8	26.7	31.3	4.7	22.9	14.6	16.0	18.9	25.3	13.5	11.2	6.4
93-1646 percent diff	0.2	1.1	9.6	9.7	10.7	5.1	11.6	3.6	11.0	1.8	7.9	5.8	12.4	9.7	0.4	8.3
94-1646 percent diff	0.2	1.9	12.2	12.0	5.8	1.0	11.6	2.2	8.0	9.1	3.2	9.2	6.6	3.4	1.2	13.2
95-1646 percent diff	0.2	2.8	12.3	12.2	1.2	6.5	13.2	6.1	5.5	0.4	4.4	5.3	11.4	8.3	8.0	13.0
AVERAGE	11.4	5.9	12.9	7.4	12.5	12.3	19	2	13.6	5.5.	9.6	11.4	16.8	8.6	4.8	9.0
91-2704 percent diff	NA	0.9	NA	20.1	21.6	21.1	28.3	0.6	25.4	17.4	19.2	19.9	22.4	17.3	17.1	5.4
92-2704 percent diff	NA	4.1	NA	7.3	7.2	9.2	11.2	15.0	9.5	0.0	5.3	5.1	8.5	1.9	2.7	9.6
93-2704 percent diff	NA	3.2	NA	8.2	5.1	7.0	13.0	0.5	12.5	9.7	8.9	8.9	10.3	6.7	6.5	10.2
94-2704 percent diff	NA	7.2	NA	5.9	4.7	1.8	12.5	0.1	9.1	4.4	4.7	10.2	6.9	3.8	3.9	21.7
95-2704 percent diff	NA	9.6	NA	4.1	3.4	5.8	9.1	1.9	9.3	6.5	9.8	8.2	10.1	9.9	6.2	18.2
AVERAGE	NA	5.2	NA	8.5	8.4	0.6	14.8	5.3	13.1	9.7	9.4	10.0	11.6	7.3	7.3	12.2
93-1ksd-3 percent diff	0.1	5.0	14.7	13.1	12.6	7.9	29.4	0.7	12.9.	5.8	7.9	8.8	12.5	7.4	2.0	10.2
95-lksd-3 percent diff	0.1	7.0	16.8	00 00	13.0	4.3	28.3	5.8	11.4	5.5	3.7	5.1	11.5	4.3	1.0	12.6
AVERAGE	0.1	0.9	15.8	11.0	12.8	6.1	28.8	3.3	12.2	2.6	2.8	5.0	12.0	5.9	1.5	11.4
	,	(			4	4		,			ı	,				
93-stsd-2 percent diff	0.1	رب 8.	18.1	12.9	13.9	15.3	30.5	1.6	15.0	20.4	7.9	1.5	2.5	9.5	0.0	2.3
94-stsd-2 percent diff	0.1	10.1	19.4	9.5	16.8	21.4	27.0	12.2	12.9	12.3	8.1	4.4	1.3	9.1	1.2	3.6
95-stsd-2 percent diff	0.1	10.4	18.6	9.01	17.1	9.5	26.0	13.0	12.9	8.3	4.9	2.0	3.6	6.5	6.0	6.1
AVERAGE	0.1	10.1	18.7	10.9	15.9	15.4	27.8	8.9	13.6	13.7	7.0	2.7	2.5	8.3	0.7	4.0
1646 certified	1.5	10.8	21	170	10.5	18	32	92	94	138	62487	413	33501	10918	372	5096
2704certified	NA	12	NA	130	14	9.86	44.1	135	95	438	61058	414	41125	12004	558	4556
stsd-2 certified	5.2	16	37	400	19	47	53	116	101	246	85230	540	58280	18660	1060	4870
Iksd-3 certified	1.9	13	30	240	30	35	47	87	82	152	66138	089	39866	12064	1440	3330

All values are percent.

Certified values from primary certificates.

NA = Data not available.

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## 2.7.5: Data Presentation and Acceptability

All SRM analyses were evaluated in two ways. First, the difference of the concentration in ppm was calculated (Table 4);

(certified value - measured value)

and then by trueness was calculated (Table 5);

(certified value - measured value) X 100% certified value

Possible analytical contamination was evaluated on the basis of these concentrations and average of concentration of the blanks. The acceptable data limits are tabulated in Table 6.

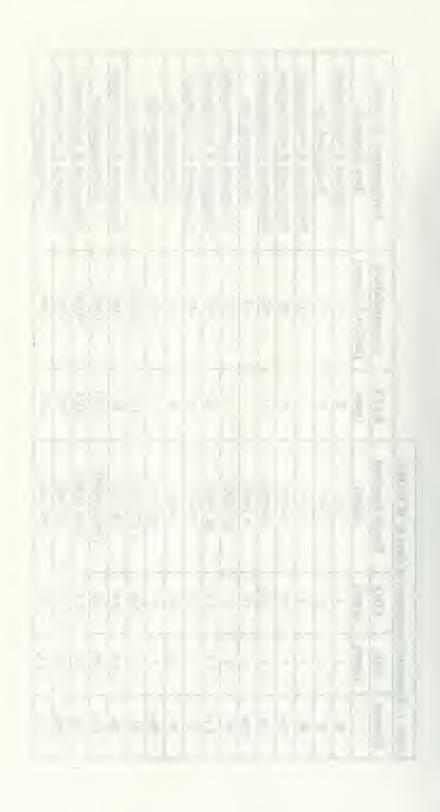
Table 6. Lowest limit (2 x HACL) for which data are acceptable.

Sc > 3 ppm	Y > 14  ppm
Sr > 100 ppm	Co > 8 ppm
Cu > 42 ppm	Ni > 32 ppm
Zn > 153 ppm	Ti > 1970 ppm
Cr > 41 ppm	V > 48  ppm
Ba > 160 ppm	Fe $> 18400 \text{ ppm}$
Mn > 190 ppm	Mg > 4150 ppm
Al > 23500 ppm	

In conclusion, for the objectives of the thesis, all data above the LOQ and/or the HACL was accepted as fit for purpose as outlined in Table 7. The effect of these limitations on the data is summarized as follows: 1) Sc, Sr, Co, Cu, Ni, Cr, V, Ba, Mg and Mn data can be used to document both background levels and anthropogenic input, as the data exceed

HACL for all elements. 2) Y data can be used to determine background levels, since they decrease in concentration in the contaminated river samples. However, anthropogenic input data cannot be assessed because the Y concentration decreases below the HACL. This surprising conclusion will be discussed in more detail in the Chapter 3. 3) Zn, Fe, Ti and Al data can be used to determine anthropogenic input but not background levels, as the data fall below HACL.

	Environments for which	Data is suitable.	Background & Anthropogenic	Background only	Background & Anthropogenic	Background & Anthropogenic	Background & Anthropogenic	Anthropogenic only	Background & Anthropogenic	Background & Anthropogenic	Background & Anthropogenic	Anthropogenic only	Anthropogenic only	Anthropogenic only	Background & Anthropogenic	Anthropogenic only	Background & Anthropogenic	Background & Anthropogenic	Anthropogenic only
	AcceptableData	(2xHACL) (ppm)	>3	>14	>100	8<	>42	>14	>32	>41	>48	>15	>150	>23500	>160	>18400	>4150	>190	>1970
	HACL	(mdd)	2	7	52	4	21	7	16	21	24	8	92	11752	82	9200	2076	95	286
assessment of data by ICP-OES.	Range Required	(mdd)	1-27	3-42	13-340	9-340	33-3900	4-1400	27-12000	48-34000	48-400	0-600	68-3100	3577-87430	89-5300	11000-260000	2288-44300	255-7800	192-11000
sessment of	L00	(mdd)	3	9 .	3	15.	15	18	9	NA	15	105	9	NA	NA	NA	NA	NA	NA
- 1	MDL	(mdd)	1	2	1	5	5	9	2	NA	5	35	2	NA	NA	AN	NA	NA	NA
Table 7: Elemental	Element		Sc	Y	Sr	Co	Cn	Mo	ï	Ç	>	W	Zn	AI	Ba	Fe	Mg	Mn	Ti



#### Chapter 3: Results and Discussion

#### 3.1: Mineralogical Analyses

The expected background mineralogy of carbonates (calcite and dolomite), and silicates (feldspars, quartz, chlorite and clays) were identified in all samples collected from the river, which are a reflection of local rock mineralogy. Unusual mineral phases present were ferrocolumbite, mullite, wuestite, with hematite and magnetite in large quantities (Fig. 10), which are unaccountable by local geology, but are minerals generally found in a number of industrial processes (Scriver et al, 1995).

### 3.2: Geochemical Analyses

Samples were extracted by two different techniques, (1) partial extraction (denoted "L") and (2) total extraction (denoted "P" and "Q"). Further total extractions were performed on samples that were separated by grain size (denoted "A" through "F" Table 8.)

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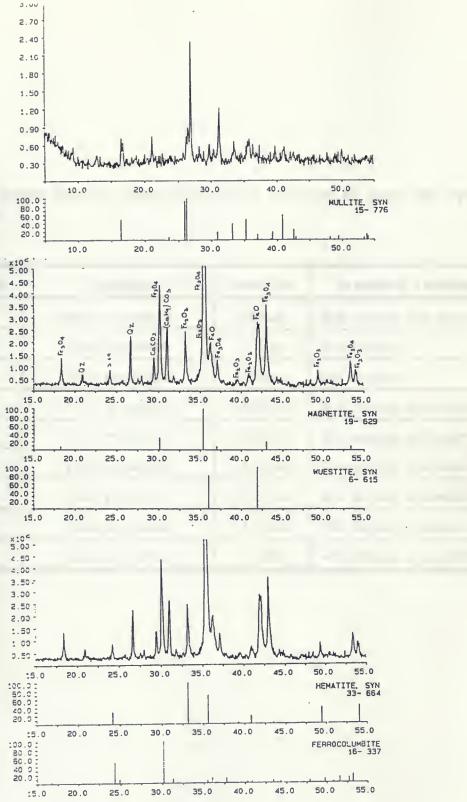


Figure 10. X-ray diffraction patterns for unusual mineral phases found at the Atlas-Mansfield discharge source.



Table 8: Sample labelling scheme according to technique for grain size separation techniques.

Symbol	Technique	Grain Size	Separation Technique
L	Partial extraction	1-63µm	Wet sieving, Dry sieving
Р	Total extraction	1-63µm	Wet sieving
Q	Total extraction	1-63µm	Dry sieving
Α	Total extraction	44-63μm	Wet sieving, cyclosizer
В	Total extraction	33-44µm	Wet sieving, cyclosizer
С	Total extraction	23-33μm	Wet sieving, cyclosizer
D	Total extraction	15-23μm	Wet sieving, cyclosizer
Е	Total extraction	11-15μm	Wet sieving, cyclosizer
F	Total extraction	<11µm	Wet sieving, cyclosizer

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## 3.2.1: Total Extraction Analyses

All total extraction analyses were evaluated to 3 sets of data (Table 9): Average continental crust (ACC) from Taylor and McLellan, 1985 (Fig. 11); the average of the control samples from the Welland River (WRC) (Fig. 12) and the lowest effect level (LEL) data of the Provincial Sediment Quality Guidelines for metals and nutrients (PSQG) (Fig. 13) from MOEE, 1993 in aquatic sediments.

Table 9.: Data used to evaluate geochemical analyses for the Welland River sediments. ACC = Average Continental Crust, WRC = Welland River Control samples, PSQG = Provincial Sediment Quality Guidelines, LEL = Lowest Effect Level, SEL = Severe Effect Level (Persaud et al., 1993).

Element	ACC(ppm)	WRC (ppm)	PSQG-LEL(ppm)	PSQG-SEL(ppm)
Be	1.5	2.07	-	-
Со	29	17.15		-
Sc	30	13.48		3-1-1-1-1
Cu	75	35.08	16	110
Ni	105	42.15	16	75
V	230	110.3	-	
Y	20	21.41	71 - 191	-
Zn	80	146.3	120	820
Sr	260	184.9	-	
Al	84100	82736	-	-
Ba	250	529.1	-	-
Fe	70700	42600	20000	40000
Mg	32000	12051	-	_
Mn	1400	380.9	460	1100
Ti	5400	5110	-	-
Cr	185	73.2	26	110

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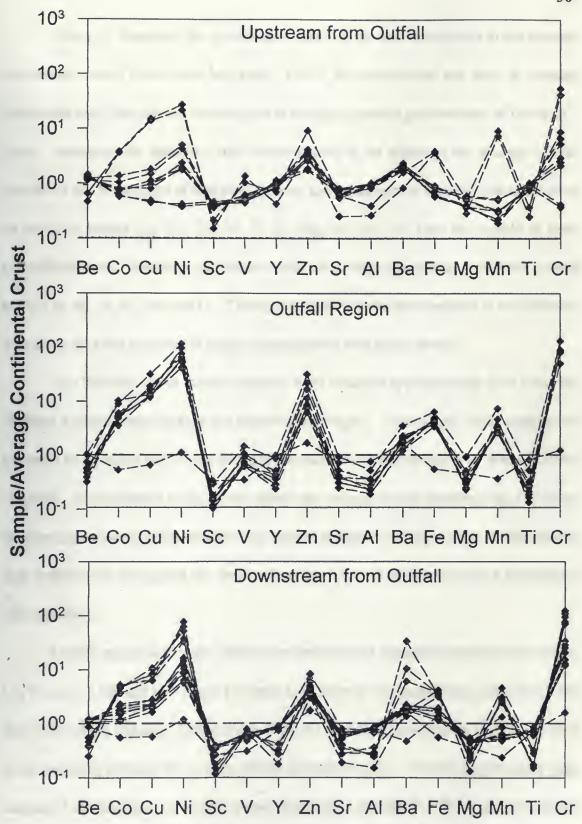


Figure 11. Geochemical data normalized to average continental crust (Taylor and McLellan, 1985).

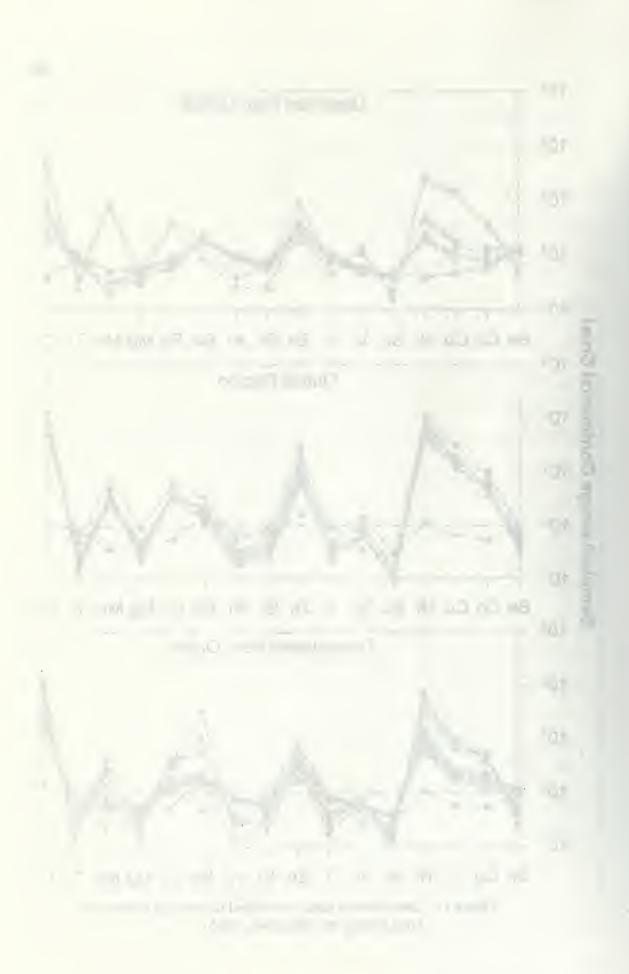


Figure 11 illustrates the enrichments found in the data normalized to the average continental crust (Taylor and McLellan, 1985). By normalizing the data to average continental crust, the data can be compared to average expected geochemistry of the earth's crust. Generally, the transition row elements tend to be higher in the average crustal abundance due to inclusion of data from igneous and metamorphic material that could have an influx of metals (eg. Co, Cu, Ni, V, Fe, Mg, Mn and Cr) from the mantle or from mineralized zones. Elemental enrichments over the average continental crust were present in Co, Cu, Ni, Zn, Fe, Mn and Cr. These particular elements are observed in the sediment indicating the river sediment is highly contaminated with heavy metals.

The Welland River control samples were obtained approximately 2km (near the Welland Airfield) upstream from the industrialized region. Accordingly, these samples are assumed to represent the natural background elemental concentrations of Welland River sediment. Normalization of the sample data to the average control samples (Fig. 12) shows the elemental concentrations relative to the control samples. The appearance of enrichments and deficiencies throughout the data set become apparent with respect to a background concentrations.

Outfall region sediments exhibit clear evidence of increased concentrations of Co, Cu, Ni, Zn, Fe, Mn and Cr. These elements are known to be contaminants associated with the steel-making industry. Downstream from the outfall, these elements are still enriched in the sediment, although the concentrations declined slightly. Three samples had a large amount of oil clinging to the sediment upon sampling (Appendix 5). These samples located approximately 50 to 100m downstream from the outfall show an anomalous increase in

barium concentrations.

Normally, it would be expected that the samples upstream from the outfall should be relatively uncontaminated. However, there is evidence of minor contamination of heavy metals. This suggests that either there could be other source(s) of contamination via other storm sewers located upstream or the river dynamics are such that contaminants move upstream from the Atlas-Mansfield outfall towards the Welland Canal siphons.

Finally, the data was compared to the Provincial Sediment Quality Guidelines (PSQG) set by the Ministry of the Environment and Energy (MOEE) in 1993. The purpose of these guidelines is to protect and manage sediment quality in Ontario waters. It is of concern that many toxic substances exist for long periods of time and can accumulate in the benthic communities (Biomagnification). Subsequently, these contaminants can transfer up the food chain (Persaud et al., 1993). The guidelines are defined to three categories: 1) No Effect Level (NEL); 2) Lowest Effect Level (LEL) and 3) Severe Effect Level (SEL) (Table 10). The criteria for these levels are summarized in the following table as outlined by the PSQG set by the MOEE.

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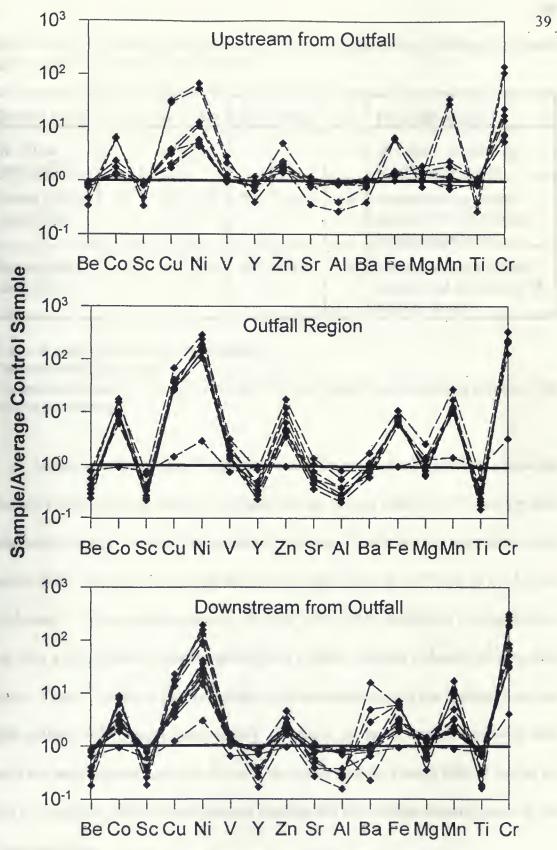


Figure 12. Geochemical data normalized to the average control samples of the Welland River.





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Table 10: Summary of data, quality levels and impact according to the Provincial Sediment Quality Guidelines (Persaud et al., 1993).

Quality Level	Cu	Ni	Zn	Fe	Mn	Cr	Potential Impact*
No Effect Level (NEL)	-	-	-	•	•	-	no impact on water or benthic community
Lowest Effect Level (LEL)	16	16	120	2%	460	26	contamination can be tolerated by majority of benthic organisms
Severe Effect Level (SEL)	110	75	820	4%	1100	110	concentrations are likely detrimental to majority of benthic species

All data in ppm, unless otherwise indicated.

According to the PSQG, if background values for metals are derived from a areas not affected by industrial point sources are greater than the "Lowest Effect Level", then the local background chemistry will form the practical lower limit. Results from the upstream control samples of the Welland River exceed the Lowest Effect Level of the PSQG for Cu, Ni, Zn, Fe, Mn and Cr. Nevertheless, comparing the data to the PSQG establishes a comparison of data with a government regulating guideline to support evidence of highly contaminated regions. Figure 13 indicates that the outfall and downstream areas of the Welland River are highly polluted with respect to heavy metals. Generally, the levels upstream are lower than that of the outfall region, although they still are higher than the Lowest Effect Level of the PSQG. In addition, there are two upstream samples that have similar contamination to that of the outfall area.

<sup>&</sup>quot;-" denotes insufficient data.

<sup>\*</sup>The potential impact is based on ecotoxic effects of chronic and long-term effects on the benthic community.

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				-	1000



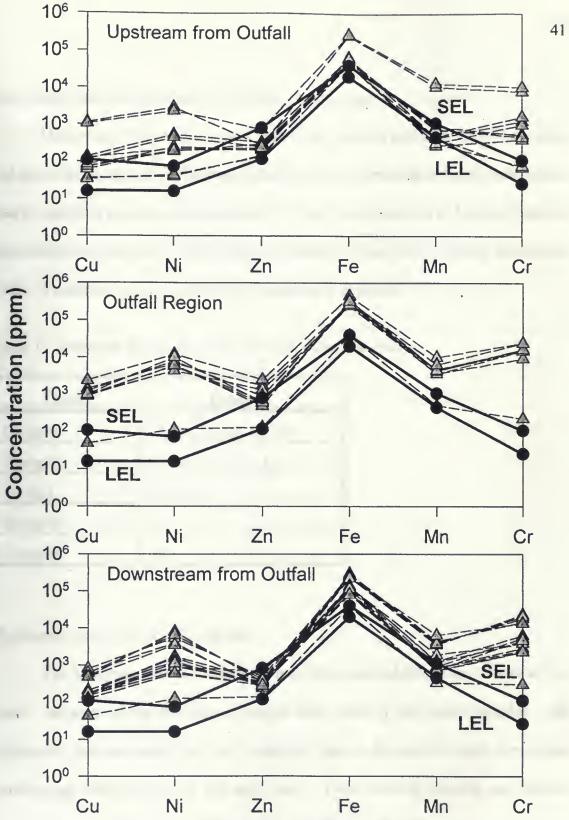
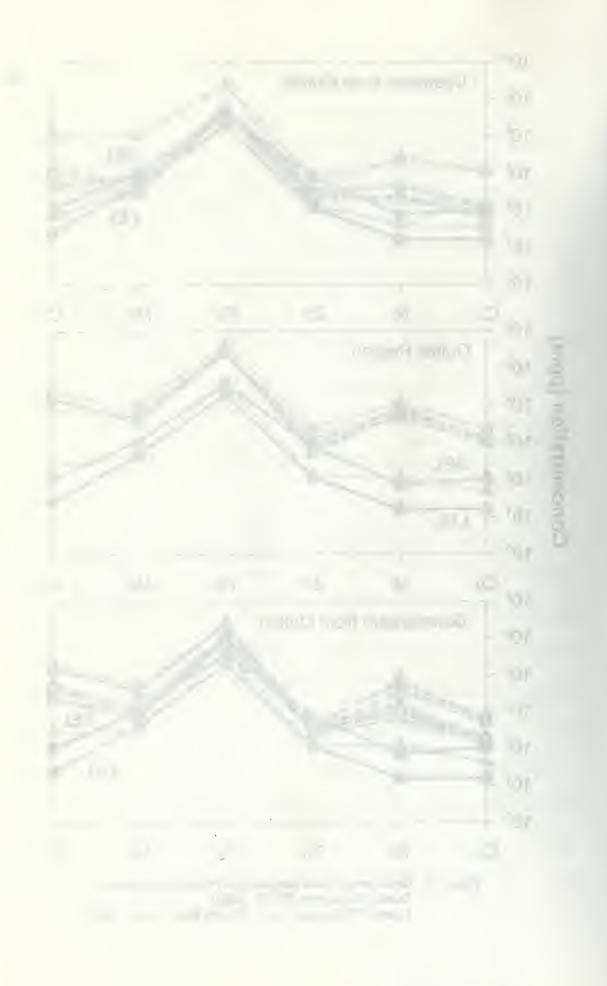


Figure 13. Geochemical data compared to Provincial Sediment Quality Guidelines (MOEE, 1993). Lowest Effect Level = LEL, Severe Effect Level = SEL.



# 3.2.2: Partial and Total Extraction Analyses - Comparisons

Metal concentrations show clear separations between upstream, the outfall region and downstream samples. Outfall region samples contain considerably higher mean values than the upstream and downstream samples. Elements were separated in 5 groups based on their pattern from upstream to downstream, elemental behaviour, and analytical differences. Table 11 outlines the groups into which each element is divided.

Table 11: Elemental groups for partial and total extraction results.

Group	Elements		
Group A	Co, Cu, Ni, Zn, Cr, Mn		
Group B	Be, Sc, Y, Sr, Ti, Al.		
Group C	V, Mg		
Group D	Fe		
Group E	Ва		

## a) Group A: Co, Cu, Ni, Zn, Cr and Mn.

The background concentrations are low, with a marked difference, a factor of 2 or more, between partial and total extraction data shown in the outfall samples. The differences between partial and total extraction data in the upstream and downstream samples are slight and exhibit the same trend. These divalent elements are first-row transition elements and are metals typically used in the steel industry.

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Total Co values approach 300 ppm, Cu and Zn concentrations are an order of magnitude higher (3000 ppm) and Ni and Mn values exceed 10000 ppm in the outfall region. The total Cr levels approach 25000 ppm and show a distinct elevation over the partial extraction data in the outfall region (Fig. 14a,b). The apparent difference in concentration of Cr between the partial and total extraction data in the outfall area suggests that the chromium resides in the total extraction phase and therefore is not immediately available to the environment. The presence of enrichments occurring in the partial extraction data, in the outfall region samples, indicate that these elements are available to the environment if the ambient environmental conditions persist for extraction (i.e. decrease in pH, increasing the acidity of the water).

## b) Group B: Be, Sc, Y, Sr, Ti and Al.

Generally, the partial extraction concentrations are low, with a relatively flat trend throughout the river samples (except Y). The total extraction values are elevated at the outfall region and have a variable trend. Be, Sc, and Y show a slight enhancement in the total extraction concentration over the partial extraction data (a few ppm). Sr shows an enrichment of 20 to 150 ppm over the partial extraction data. Al and Ti show marked differences between the extraction results (10000 - 65000 ppm and 600 - 5000 ppm, respectively) (Fig. 15a,b).

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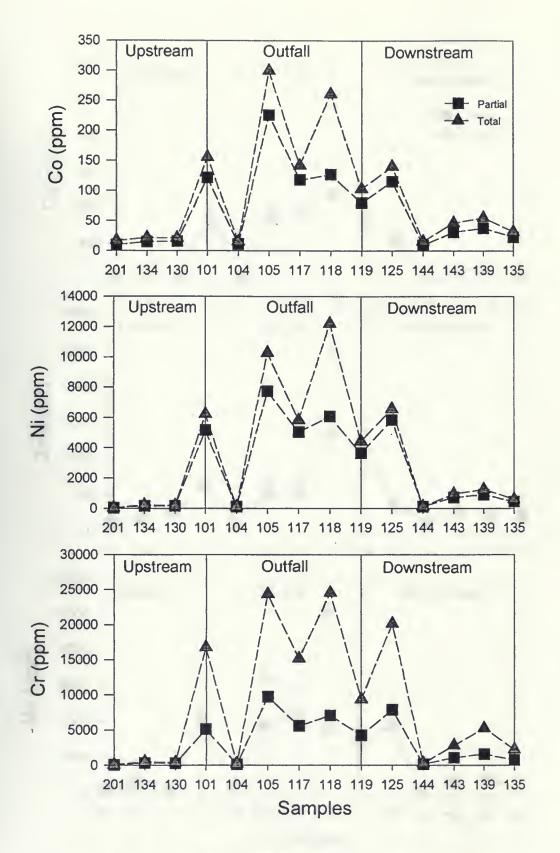
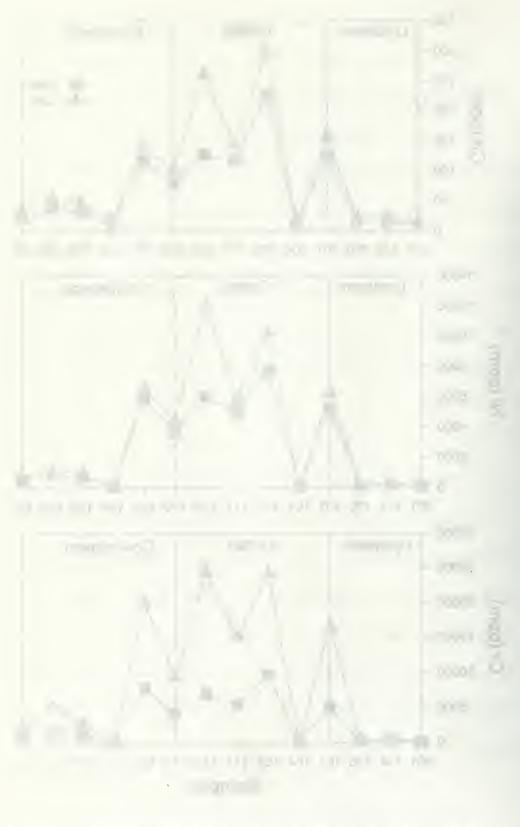


Figure 14a. Partial and Total extraction comparisons for Group A.



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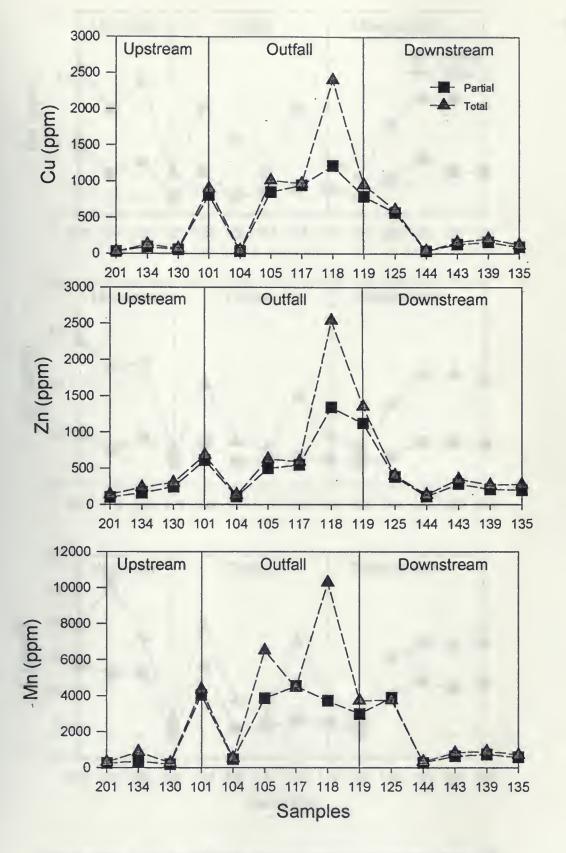
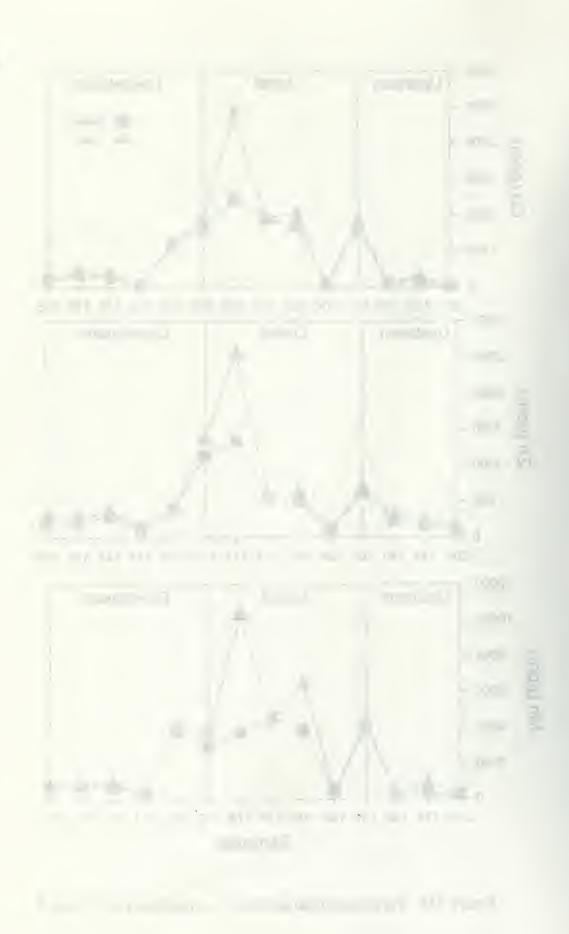


Figure 14b. Partial and total extraction comparisons for Group A.



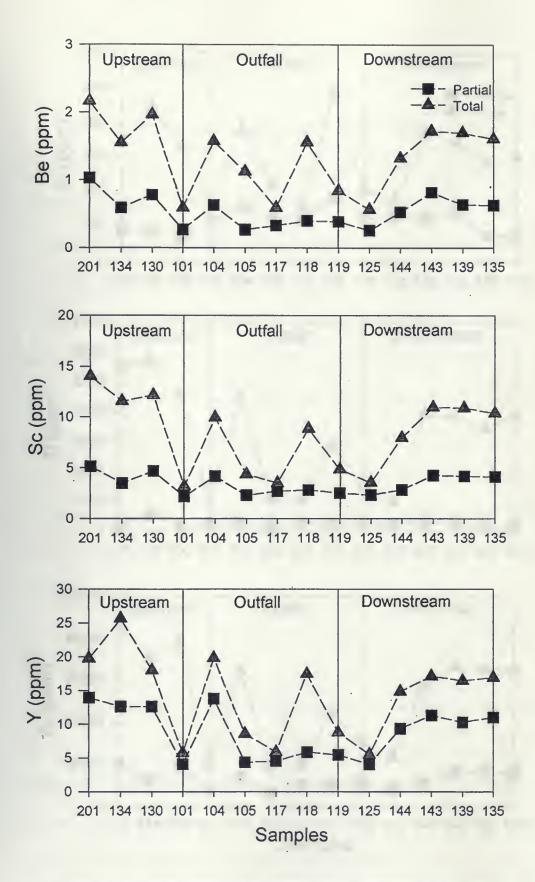


Figure 15a. Partial and total extraction comparisons for Group B.



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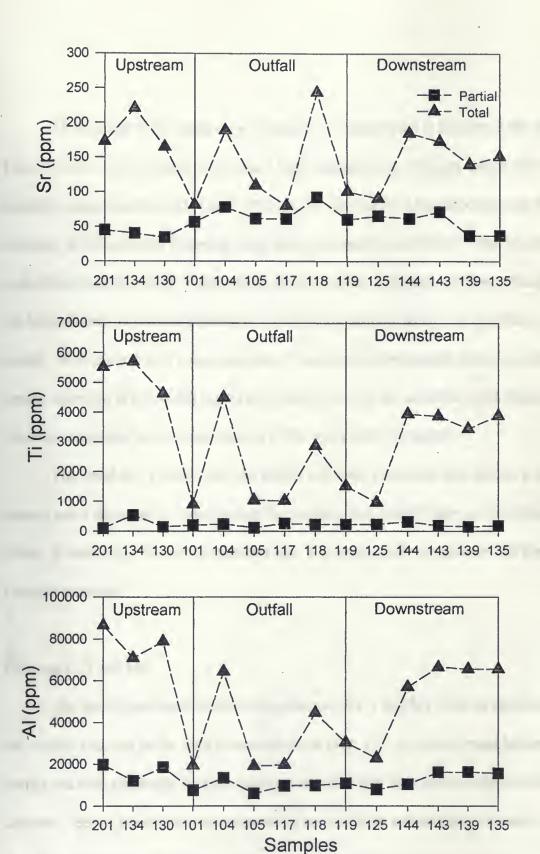
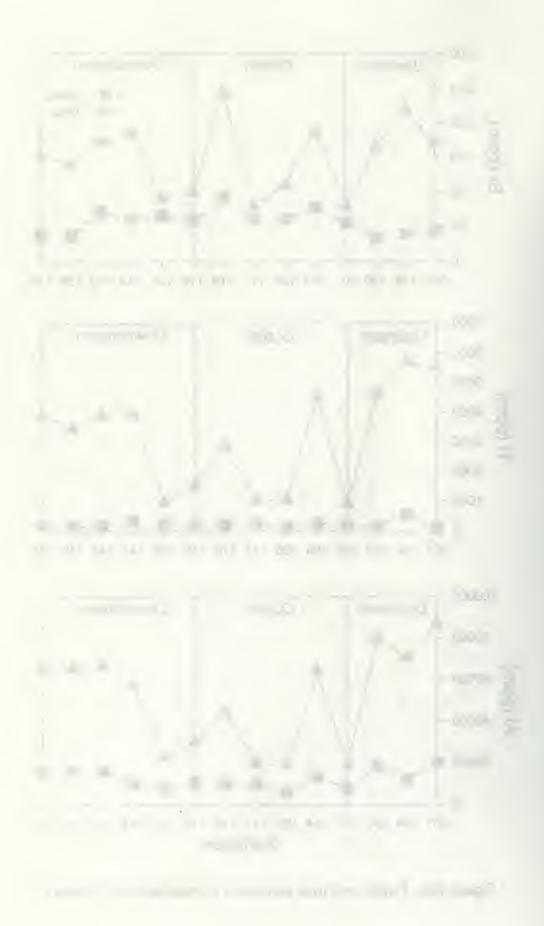


Figure 15b. Partial and total extraction comparisons for Group B.



All elements of this group show a decrease in concentration in samples at the outfall. These samples in the outfall area have a high concentration of heavy metals (Group A elements) suggesting that the Group B elements are derived from the carbonate and silicate structure of the minerals occurring there (compositionally controlled). This decrease in concentration in the outfall region and an increase in the upstream and downstream areas can be attributed to a dilution effect of mill scale anthropogenic sediments introduced at the outfall. With the influx of a large amount of water and anthropogenic particles, often the natural sediments in the outfall region are washed away by the turbulence and diluted with the newly deposited particles settling out of the water from the outfall.

The trend for Y show both the partial and total extraction data follow a similar pattern and a decrease in concentration for samples that usually show an enrichment of Group A elements. This trend indicates that Y is more easily extractable that the other Group B elements.

## c) Group C: V and Mg.

The partial and total extraction concentrations for V and Mg show an enrichment in the outfall area and in the total extraction phase (Fig. 16). A similar trend between the partial and total extraction for both elements indicates that they are extractable with each digestion. Hence, these elements are probably derived from anthropogenic mineral phases and/or natural minerals (e.g. iron oxides and carbonates).

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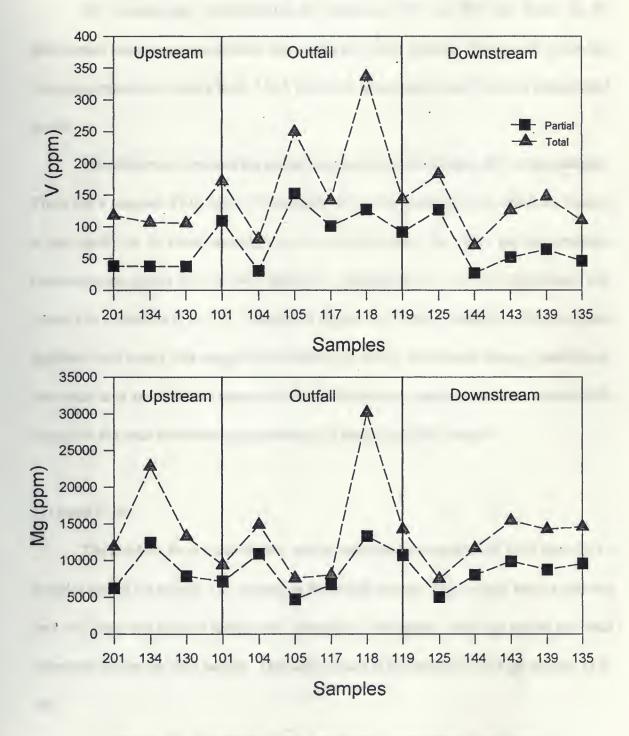
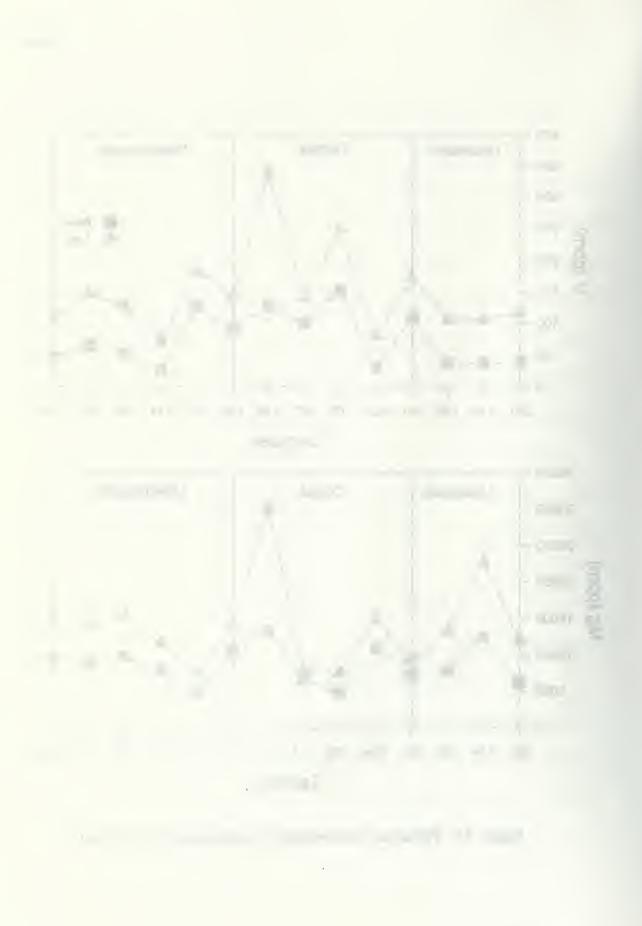


Figure 16. Partial and total extraction comparisons for Group C.



#### d) Group D: Fe.

The background concentrations are relatively low and the iron levels in the downstream samples are considerably lower than the outfall samples. The outfall region has iron concentrations ranging from 3 to 5 orders of magnitude higher than the background levels.

The differences between the partial and total extraction results for Fe are complex. There are 4 samples, (101, 105, 117 (outfall), 125 (downstream)) all of which are located at the outfall or in close proximity to the outfall region, that have partial extraction concentrations greater than the total extraction concentrations. Fe is the only element that shows this behaviour (Fig. 17). Incomplete digestion of heavy minerals or anthropogenic particles (mill scale), like magnetite and hematite, which were found in large quantities at the outfall area are likely the reason the partial extraction concentrations are elevated with respect to the total extraction concentrations in these particular samples.

## e) Group E: Ba.

The trend for Ba is relatively flat, with a concentration variation of 1000 ppm for all samples except for sample 125, located in the outfall region. This sample has a value two and half times the relative background chemistry (2500 ppm). Both the partial and total extraction follow the same pattern. This trend occurs in the anomalously high sample (Fig. 18).

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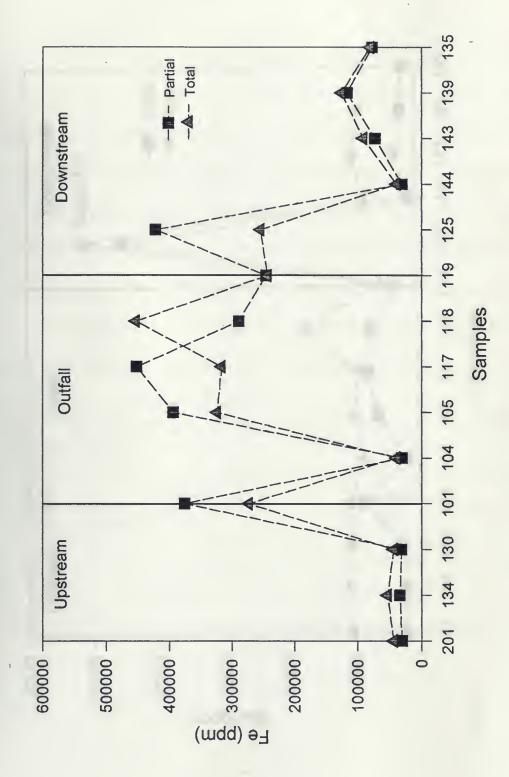


Figure 17. Partial and total extraction comparisons for Group D.



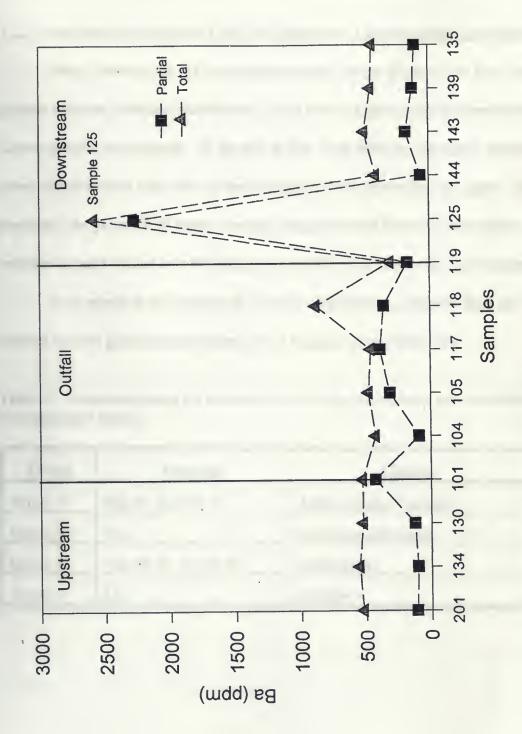


Figure 18. Partial and total extraction comparisons for Group E.



## 3.2.3: Geochemical Analyses of Grain Size Separates - Concentration Comparison

Metal contents are not homogeneous over a range of grain size fractions. Finer grained fractions, containing mainly clays, often show a higher metal concentration than its coarse grained counterparts. In the silt to fine sand fraction, the metal concentrations generally decreased with size as the fraction is more dominated by quartz. In coarse fractions, the presence of heavy minerals (magnetite and hematite) may cause a further increase in metal content over the background geochemistry (Salomons and Forstner, 1984).

In the results of the grain size to chemistry correlations, generally there are 3 groups defined by their geochemical patterns across the grain sizes (Table 12).

Table 12: Elemental groups for grain size to chemistry correlations and controlling environmental factors.

Group	Elements	Control	
Group 1	Co, Ni, Mn, Fe, V	Anthropogenic (metals)	
Group 1b	Ba	Anthropogenic (oils)	
Group 2	Mg, Sr, Ti, Al, Y, Sc	Lithological	
Group 3	Cu	Unknown	

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Group 1: Co, Ni, Mn, Fe, V.

Group 1 is a suite of elements that exhibit no significant correlation with grain size (Fig. 19a,b). In typical sediments, the enrichment in concentration tends to be in the finest fraction. This occurs due to several physical, chemical and mechanical properties such as; large surface area on clay size sediments, cation exchange capacity, magnetic properties, surface charge and composition. This lack of the expected increasing concentration toward the fine grained sediment pattern is the function of the input of anthropogenic mill scale particles into the environment. With increased levels of heavy minerals (hematite and magnetite), often the grain size pattern appears flat, with the occasional peak or anomaly. For example, the peak that is consistent at grain size fraction D, occurs for Co, Ni, Mn and V. These anomalies represent field sample 101, which is located directly at the outfall. The Fe and Mn patterns show an increase in concentration at both A and F. That increase should normally be present at F due to physical and chemical partitioning. The presence of these heavy minerals suggests that source of Fe and Mn contamination or at least in part is derived from the particulate load. The majority of the Fe concentrations range from 1% to 15%, with the exception of three samples (125, 124 and 101). Sample 101 is located at the outfall, the remaining anomalies are located less than 100m downstream from the outfall. The concentrations of Fe in these samples ranges from 15% to 25%, therefore strengthens the argument that the iron is in the form of anthropogenic mill scale grains and is derived from the steel-making industry.

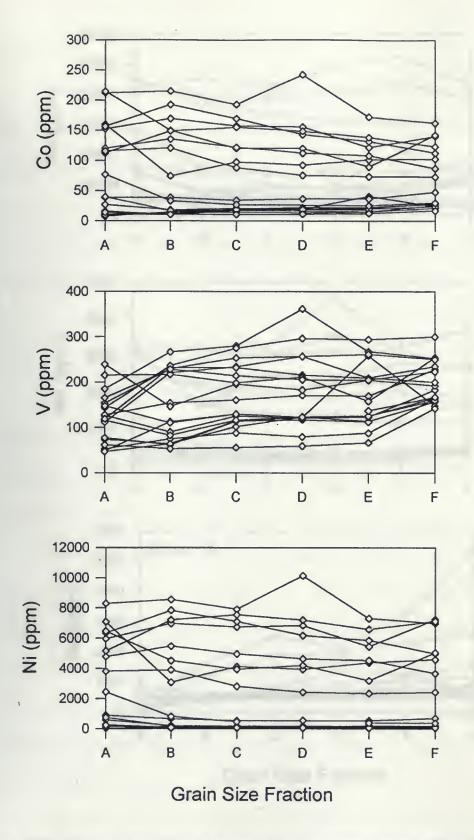
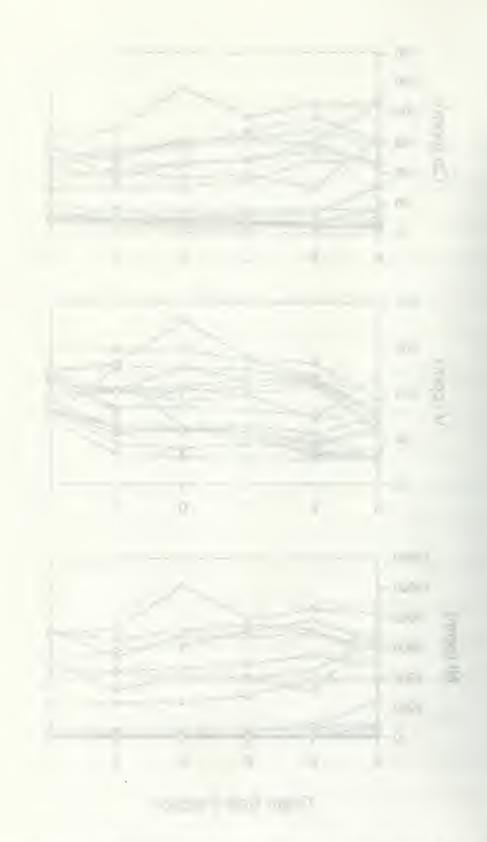


Figure 19a: Grain size to chemistry correlation comparisons for Group 1 and 1B. A=63-44, B=44-33, C=33-23, D=23-15, E=15-11, F=<11.



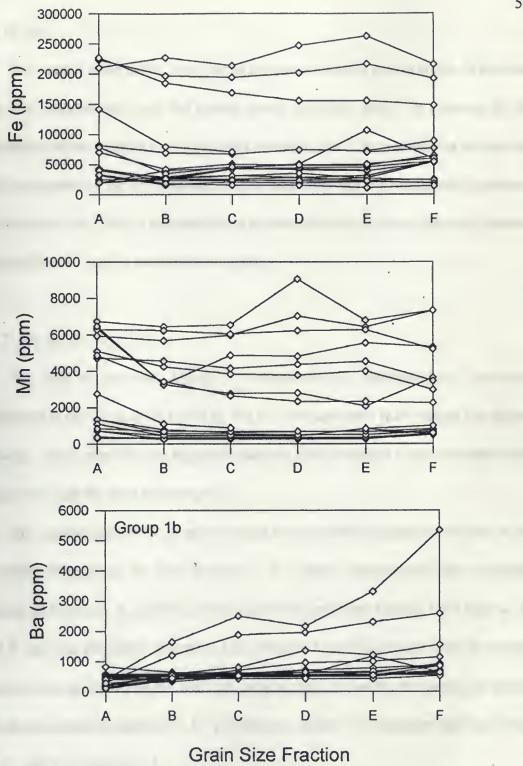
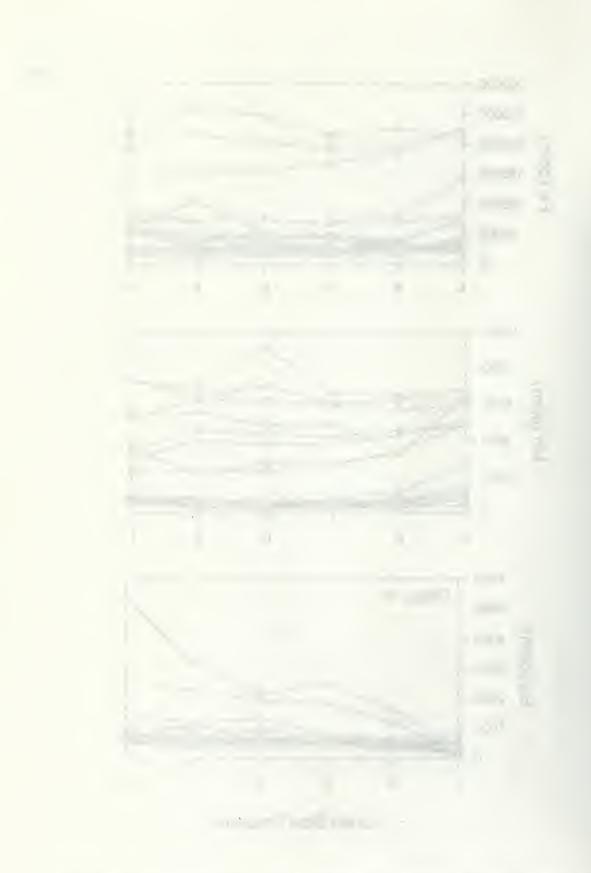


Figure 19b: Grain size to chemistry correlation comparisons for Group 1 and 1B. A=63-44, B=44-33, C=33-23, D=23-15, E=15-11, F=<11.



Group 1b: Ba.

The overall trend of Ba is similar to the other elements in this group, in that they have a low concentration and flat pattern across the grain sizes. In contrast, the Ba concentrations of two samples (125, 124) are extremely high. These particular samples are located approximately 100 m downstream from the outfall and were observed to contain a large quantity of oil. The Ba concentrations are only elevated in these individual samples implies that the Ba may be attributed to the oils.

Group 2: Mg, Sr, Ti, Al, Y, Sc.

The suite of elements (Group 2) demonstrates the relationship of increasing concentration to the finest grain size (Fig. 20a,b). The parameter that controls this pattern is lithology. These elements are major components in the minerals of the carbonate-shale basement rock and the soils in the region.

The concentrations of Sr, Y and Ti appear to be variable in grain size fraction A and less variable throughout the finer fractions. All content demonstrates this variability throughout all fractions. A positive anomaly occurs in grain size fraction E for Mg, Sc, Ti, Sr and Y and was identified as sample 132, which is located upstream from the outfall, although in close proximity to the Welland Canal siphon. However, this particular sample also produces a negative anomaly in Al. Furthermore, sample 115 demonstrates high values for Ti, Sr and Y in grain size A.

The patterns produced by Al and Ti suggest that the concentrations of these elements may be attributed to clay minerals such as illite. Since Sr levels are greatest in the coarse

fraction it is likely derived from the carbonate minerals. Furthermore, Mg values are consistent across the grain size range also supports the X-ray patterns identification of detrital dolomite in the samples. Sc and Y produce a trend expected of background elements (not influenced by metal contamination). The concentrations of these elements are directly related to the silicate and/or carbonate bond structure. Finally, Sc and Y concentration variability between samples is not significant in terms of analytical uncertainty due to the low concentrations approaching the method detection limit. However, they were included in this group because they produce a consistent pattern similar to the other elements in this group.

#### Group 3: Cu.

Cu was placed into a group by itself due to the strong correlation between grain size and chemistry (Fig. 21). The pattern shows a steady increase in copper content up to grain size E and then a slight decrease at grain size F. This trend is perhaps explained that industrial Cu is of grain size range E (11-15µm). The trend is similar to group 2, with the exception of the decrease in Cu concentration at grain size fraction F. In contrast to the partial and total extraction data, Cu behaved similarly to Co, Ni, Zn, Cr and Mn. Group 3 demonstrates the same anomalies, sample 101 at grain size D, sample 132 at grain size E, in Groups 1 and 2 respectively. In addition, sample 119 has a high Cu value (4000 ppm), a factor of 4 above background levels, at grain size E.

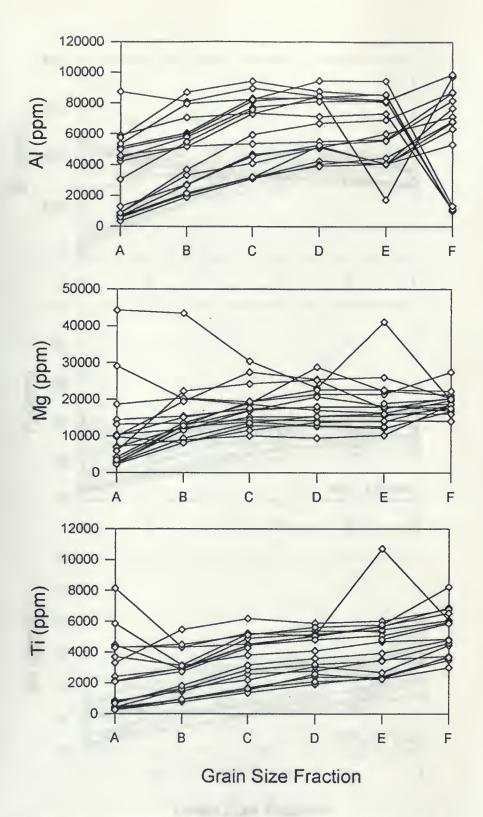


Figure 20a. Grain size to chemistry correlation comparisons for Group 2. Grain sizes as in Figure 19.



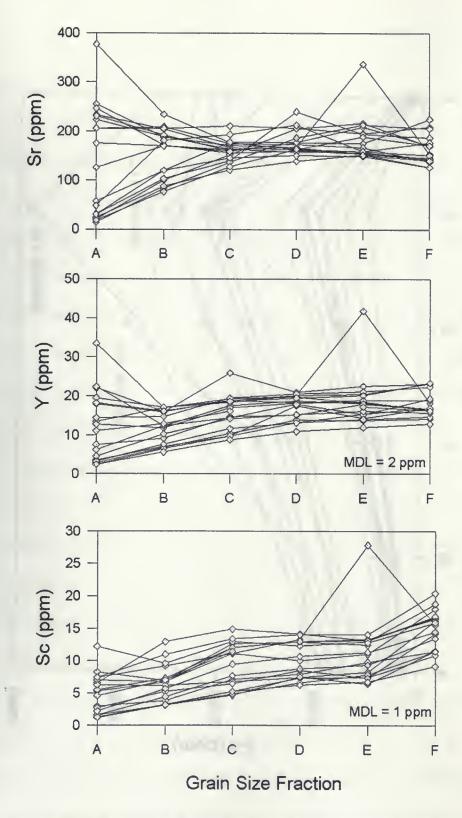


Figure 20b. Grain size to chemistry correlation comparisons for Group 2. Grain sizes as in Figure 19.



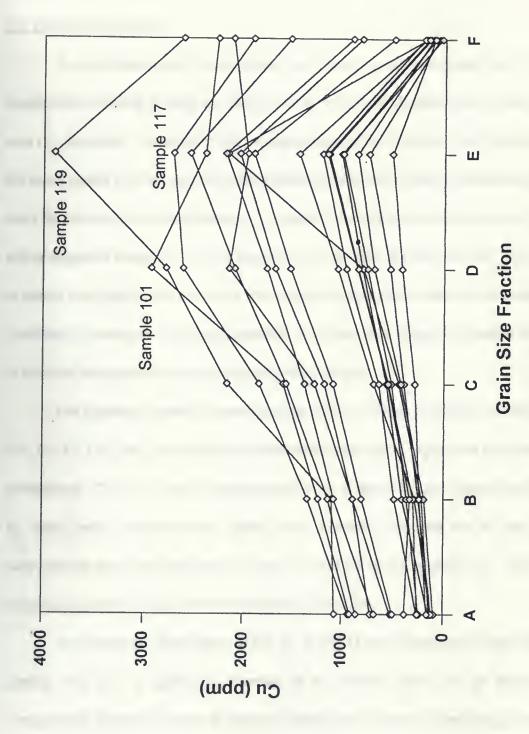


Figure 21. Grain size to chemistry correlation comparisons for Group 3. Grain sizes as in Figure 19.



## 3.3: Dispersion Patterns

River sediments are deposited and resuspended by erosion depending on the river morphology (scouring of banks and deposition on point bars), flooding and influences from man (ie. dredging). As particles are transported along the river bed, they bring with them the heavy metals that are comprise in the sediment in the form of heavy minerals or that they carry by chemically bound processes. Eventually, the sediment will settle out of solution and be deposited elsewhere. This is especially true for fine-grained material. The location at which it is finally deposited may have a greater impact than where it originated from. Therefore, assessment of the contamination levels and their dispersion pattern are crucial in terms of resuspension and consequently redeposition.

The dispersion patterns demonstrated in Figure 22 clearly indicate the heavy metals (Cu, Co, Ni, Cr, V and Zn) are highly concentrated in the outfall region and decrease further downstream. Cu, Co, V and Zn contamination levels at the outfall are approximately greater by three orders of magnitude. Metal contamination is greatest for Ni and Cr with concentration levels ranging from 0.5% to 1.5% and 1% to 3% respectively. The relative dispersion pattern between these two elements is similar.

The elemental contribution of Be, Sc, Y, Sr, Ti and Al exhibits a relatively similar pattern (Fig. 23); a significant decrease at the outfall region for all analytes. The concentration differences between upstream, outfall and downstream and are greatest for Ti and Al. The upstream distribution represents background concentrations of typical river sediment of this mineralogical composition (Chapman and Putnam, 1984). The downstream

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patterns for Be, Sc and Al return to the usual pattern. The downstream distribution for Y, Sr and Ti appears erratic. This overall distribution is inverse with respect to the pattern observed by the heavy metals in Figure 22. The decrease in concentration at the outfall in Figure 23 is attributed to a dilution of the naturally occurring sediments.

The dispersion pattern in Figure 24 are individually unique. Ba concentrations remain relatively flat and consistent throughout the river, with the exception of 3 samples that have high values. It is probable that the Ba is related to the oils observed in the river. The Mg pattern is similar to that of Figure 23, with the exception of some of the downstream samples show some depletion of Mg; suggesting a possible mixed anthropogenic (steel) and background (carbonates) source for Mg. Finally, the dispersion patterns for Fe and Mn are similar, however vary dramatically in concentration. The apparently high Fe levels observed at the outfall coincide directly with the metal filings found in the sediment and subsequently identified with XRD as magnetite, hematite and wuestite. Both the Fe and Mn can be attributed to an anthropogenic source.

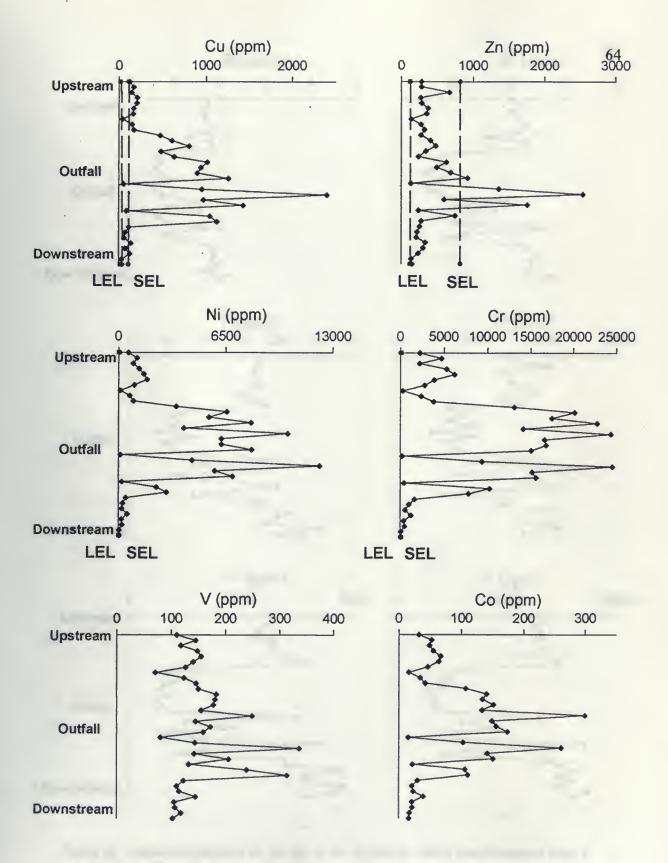


Figure 22. Dispersion patterns for total metal concentrations of Cu, Zn, Ni, Cr, V, and Co. Heavy metals are highly concentrated at the outfall region.

LEL = Lowest Effect Level, SEL = Severe Effect Level (MOEE, 1995).



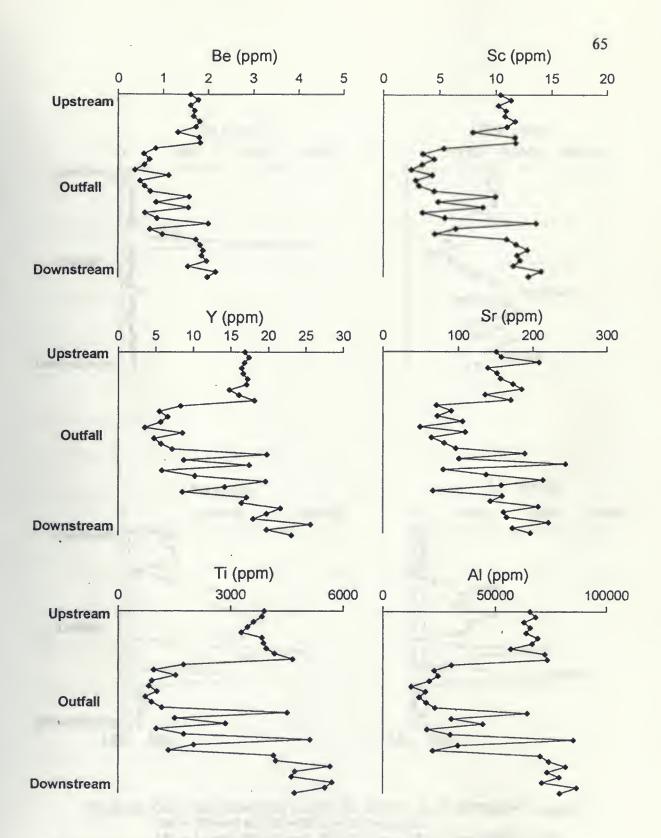
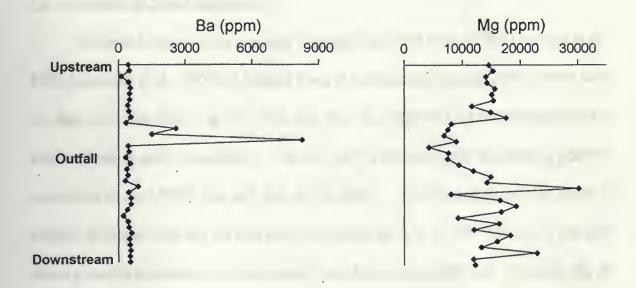


Figure 23. Dispersion patterns for Be, Sc, Y, Sr, Ti, and Al. Metal concentrations show a significant decrease at the outfall area. This decrease clearly shows the dilution effect by mill scale and other anthropogenic materials on river sediment composition.





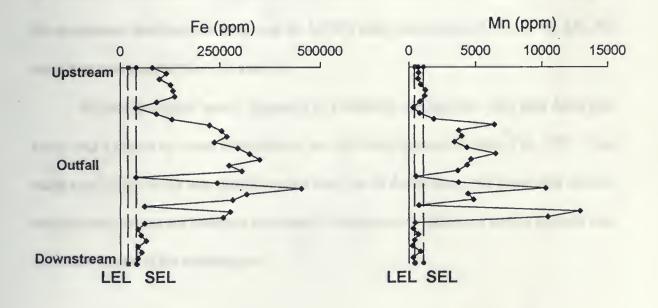
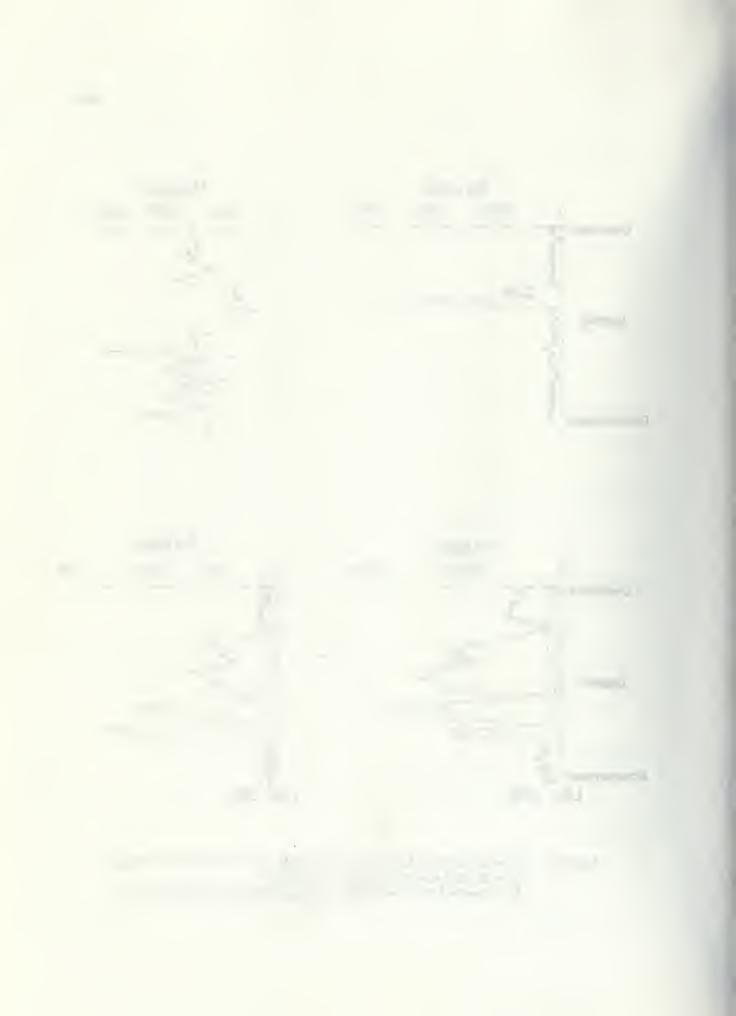


Figure 24. Dispersion patterns for Ba, Mg, Fe, and Mn. Each distribution is unque in that a different source attributes to the pattern.

LEL = Lowest Effect Level, SEL = Severe Effect Level (MOEE, 1995).



## 3.4: Discussion of Data Comparisons

Elemental comparisons in scatter diagrams have data from MOEE (Persaud et al., 1993, Jaagumagi et al., 1995) and Niagara River (Mudroch and Duncan, 1986) plotted with the data from this study (Fig. 25). The data from this study and the MOEE study have a similar trend in each comparison. The Cu and Ni chemical plot illustrates a positive correlation in the MOEE data and data of this study. The Niagara River data shows a slightly different trend and the data does not exceed the SEL for Ni. The Fe vs. Mn plot shows a positive correlation with increasing Fe and Mn in the outfall area. The majority of the Zn and Cr concentrations of the Welland River are higher that of the Niagara River. Yet, the Welland River sediment and Niagara River sediment exceed the PSQG for Cr. Welland River sediment data from this study and the MOEE study have levels of Ni, Cu, Fe, Mn, Zn and Cr exceed the PSQG's LEL and SEL.

Figure 26 depicts scatter diagrams of elemental comparisons with data from this study and a report by Acres International Inc. for Atlas Speciality Steels Co., 1991. The study area for this thesis was slightly smaller than that of Acres study, but the results of both studies conclude that the sediment was heavily contaminated with heavy metals and oils and was recommended for remediation.



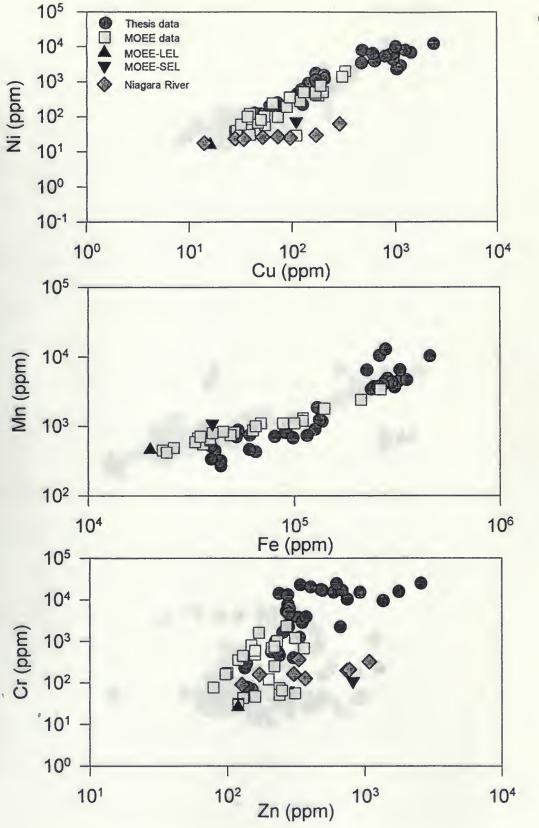
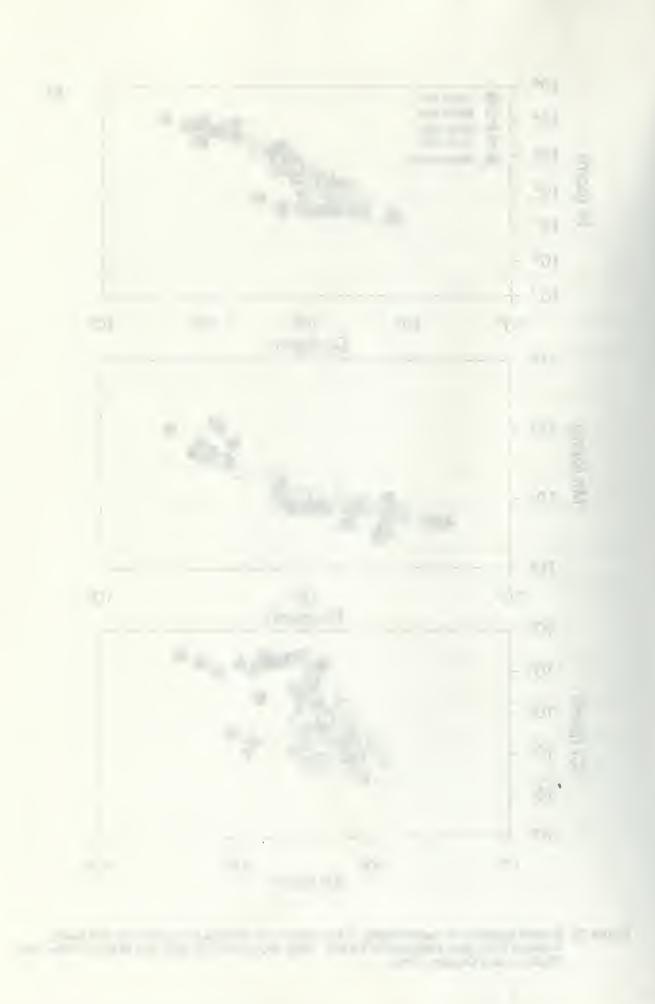


Figure 25. Scatter diagrams of geochemistry of the Welland River data produced from this study, Welland River data produced by MOEE, 1995, MOEE PSQG, 1993 and Niagara River data (Mudroch and Duncan, 1986).



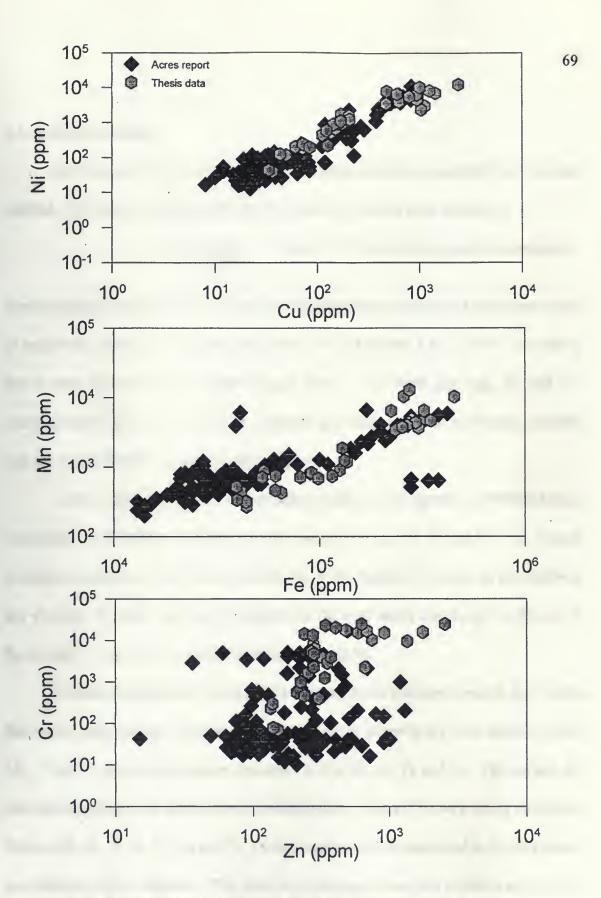
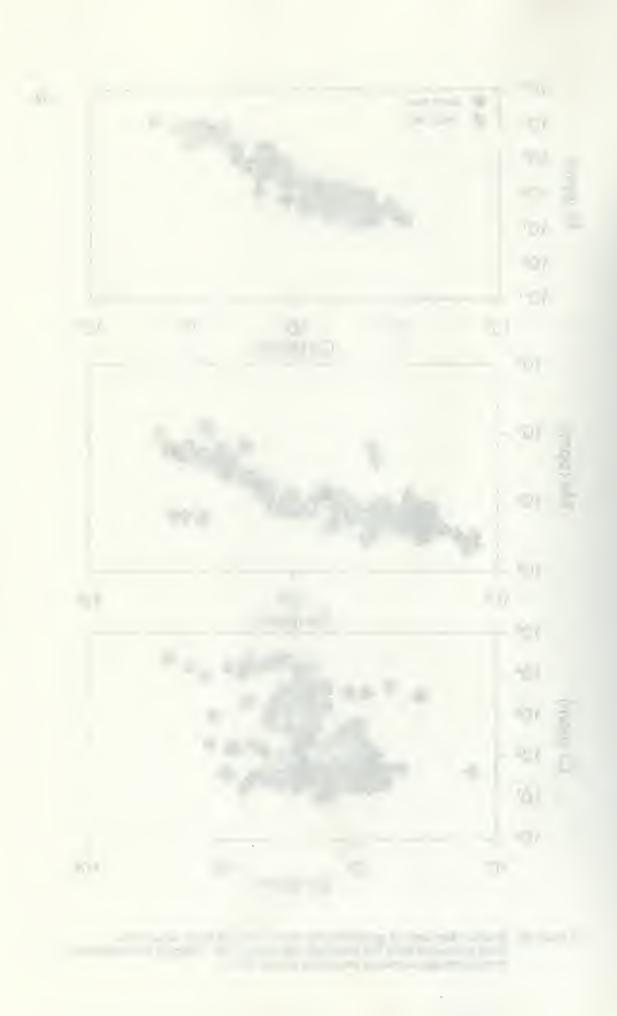


Figure 26. Scatter diagrams of geochemistry of the Welland River sediments.

Data produced from this study and Welland River dredging demonstration environmental workshop (report by Acres, 1991).



## 3.5: Statistical Analysis

Statistical analysis was performed on elements derived through the total digestion method. The data was standardized by the following formula prior to analysis.

$$Z = \frac{(X_1 \times X)}{STD}$$
, where  $X_1$  is the actual measured concentration

Standardizing the data, makes each observation independent of units and significate orders of magnitude (unitless) of the same unit (Davis, 1973, Marriott, F.H.C., 1974). In order to assess data that have concentration ranges from 1 to 100000 ppm (eg., Sc and Fe), standardization allows the data to be compared as similar units and to compare variables that are very different (e.g. apples and oranges).

Table 13 reports a matrix of correlation coefficient for variables of Welland River sediments. An excellent correlation exists among Co, Cu, Ni, Fe and Cr. An inverse excellent correlation is present between Be, Sc, Y, Sr, Al and Ti. Ba has no correlation to any element. A weak correlation is apparent in Zn to all heavy metals, but not Be, Sc, Y, Sr, Al and Ti. Mg shows a good correlation to Y and Sr.

Varimax rotated factor analysis of the Welland River sediments records four factors that explain geochemical variation between loadings of elements for these samples (Table 14). Factor 1 shows a correlation between Co, Cu, Ni, Zn, Fe and Cr. This reflects the elements associated with anthropogenic contamination. Factor 2 shows a strong correlation between Be, Sc, Y, Sr, Al, Mg and Ti. These elements are with associated to the background geochemistry of the sediment. The close, high loadings among the elements accounts for

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a distribution of elements typical of sediment composition. The third factor shows a strong correlation between V and Mn and accounts for 15% of the variation. These elemental levels are contributed from both background and contaminate chemistries. Subsequently, these 3 factors also support the correlation found in the Pearson correlation matrix and account for 85% of the total variation.

The correlation matrix and factor analysis further support the observations and results presented earlier in this chapter. The heavy metals (Co, Cu, Ni, Fe, Mn and Cr) exhibit strong correlations among each other and an inverse correlation to the background elements (Be, Sc, Y, Sr, Al, Mg and Ti).

Table 13: Correlation Coefficient Matrix among Chemical Analyses for Welland River Sediments

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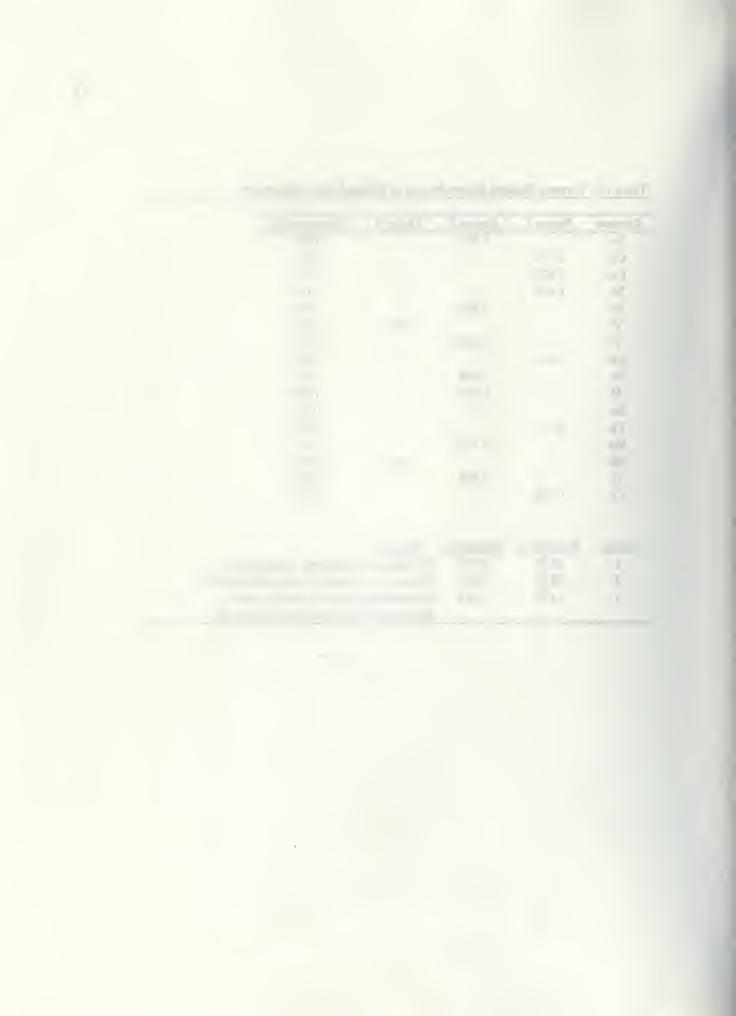
Τ̈́														-	-0.862
Mn													_	-0.656	0.675
Mg												_	-0.06	0.597	-0.311
Fe											-	-0.217	0.81	-0.864	0.925
Ва										_	0.152	-0.185	0.042	-0.272	0.413.
Al									_	-0.258	-0.875	0.558	-0.697	0.974	-0.865
Sr								-	0.789	-0.155	-0.562	0.829	-0.368	0.825	-0.564
Zn							1	0.072							
Y						-		0.898							
>					1										0.685
Sc				-				0.805							
Z			1	-0.745											
Cu		1	0.87												
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Be (				'				'						1	
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# NUMBER OF OBSERVATIONS: 35

Table 14: Varimax Rotated Factor Matrix of Welland River Sediments

Element	Factor 1	Factor 2	Factor 3	Communality
Be	-	0.815	-	1.000
Co	0.793	-	-	1.000
Cu	0.823	-	-	1.000
Ni	0.856	-	-	1.000
Sc	-	0.830	-	1.000
V	-	-	0.803	0.975
Y	-	0.906	-	0.992
Zn	0.879	-	-	0.953
Sr	-	0.948	-	0.95
Al	-	0.791	-	1.000
Ba	-	-	-	0.879
Fe	0.731	-	-	0.995
Mg	-	0.892	-	0.941
Mn	-	-	0.809	0.954
Ti	-	0.818	-	0.991
Cr	0.718	-	-	0.995

Factor	% Variance	Eigenvalue	Diagnosis
1	32.57	10.567	Elements associated with contamination
2	38.31	2.915	Elements associated to background chemistry
3	14.55	0.534	Elemental levels contributed from both
			background and contaminate chemistries.



# **Chapter 4: Summary and Conclusions**

The investigation of a site of the Welland River was initiated to describe chemical dispersion patterns in close proximity to the outfall region as compared to samples obtained upstream from the industrialized zone.

The results of this study can be used as a template in terms of analytical method development and modification. Numerous analytical techniques, typically used on soils, sediments, and rocks were applied. Some were found to work well, while other techniques did not. Generally, the typical sample pretreatment and separation techniques for sediments worked well. Good laboratory practices were used to ensure integrity of the samples prior to chemical analysis. In contrast, the usual extraction and digestion techniques did not work well. Due to the complex nature of the samples, much research, development, and modification had to be employed in order to ensure the most accurate and precise data possible. The amount of iron oxide minerals found in the samples, those in particular, found at the outfall, were responsible for difficulty in dissolving the sediment.

The mineralogy of the Welland River sediments contains expected background minerals (calcite, dolomite, quartz and clays). These mineral phases are directly related to the background carbonate-shale basement and the glacial clay plains found in the Niagara Peninsula. Others minerals that cannot be attributed to the local geology (eg. mullite and ferrocolumbite), are likely associated with the industrial companies in Welland.

Normalization of total digestion data exhibits clear evidence of heavy metal contamination in comparison to the average control samples obtained upstream from the effected area. Comparison of the results to average continental crust (Taylor and McLellan,

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1985) also suggests heavy metals are elevated in the study area. Chemical analyses of the sediments were compared to the Provincial Sediment Quality Guidelines (PSQG) devised by the Ontario Ministry of Environment and Energy (MOEE), since the MOEE has been involved with Atlas Specialty Steels to eliminate or minimize the impact in Welland. These results concluded that metal concentrations (Cu, Ni, Zn, Fe, Mn and Cr) are in excess of the PSQG severe effect level (SEL) at the outfall region. In addition, samples upstream and downstream from the outfall region contain metal concentrations exceeding the PSQG lowest effect level (LEL).

Partial and total extraction results were divided into 5 groups based on their chemical distribution, elemental behaviour and true analytical differences. The patterns observed in the partial extraction data tends to dependant upon the extractability of the element or phase that binds the analyte. This in turn, determines the ease of availability of metals into the surrounding environment and, therefore, its impact on the organisms that depend on that environment. The majority of analytes were lower in concentration in the partial extraction relative to the total digestion results, with the exception of iron. This suggests that most of the elements have an acid extractable phase. Therefore, these trace elements are bound in exchangeable, carbonate, oxide, or organic phases. The total extraction values are generally higher than the partial extraction concentrations, suggesting that the acid attack use in takes most analytes bound to exchangeable, organic, carbonate, oxide and silicate phases. However, it was observed that the partial extraction concentrations were greater than the total extraction concentrations for iron for four samples located at the outfall region. It is suggested that incomplete digestion of heavy minerals or anthropogenic particles (magnetite

and hematite) resulted in this abnormal behaviour between partial and total extraction for iron.

Geochemical analyses of six grain size separates were divided into 3 categories based on their geochemical patterns of grain size separates from fine sand (63μm) to clay (4μm). The correlation observed for group 1 elements were identified as a function of an anthropogenic input in to the system. Group 2 elements demonstrate an increasing concentration to the finest grain size suggesting this trend is directly related to the silicate and/or carbonate mineralogy. Finally, Cu was designated to its own group due to the strong correlation between grain size and chemistry. According to several researchers observations (Salomons and Forstner, 1984, and Forstner, 1982), the fine grained material should possess the highest concentrations of metals. In contrast, the pattern observed in copper is slightly different than expected. The concentrations increase toward the finer grained material, before decreasing significantly at the finest grain size. It is unknown what controls this result.

The dispersion patterns are demonstrated largely by two distinct categories 1) the heavy metals (Co, Cu, Ni, Zn, V and Cr) and 2) the background elements (Be, Sc, Y, Sr, Al and Ti). The heavy metals show a significant increase at the outfall region and decrease further downstream. Ni and Cr levels are extremely high at the outfall. Elemental concentrations of background elements exhibit a marked decrease at the outfall, then returning to higher levels downstream. This negative anomaly observed in these elements at the outfall is attributed to a dilution effect of the natural sediment and subsequent concentration of mill scale sediment. Generally it would be expected that consistent



environmental conditions would exist locally. However, it was recognized that "hotspots" or "non-hotspots" occur in the river. Sample 104 is located at the outfall region, adjacent to samples that were highly contaminated with heavy metals, yet the chemistry of this particular samples resembles that of the controls samples.

Multivariant statistical analysis and correlation coefficient matrix results concluded that the heavy metals (Co, Cu, Ni, Zn, Fe, Mn and Cr) correlate strongly. Accordingly, the background elements (Be, Sc, Y, Sr, Al and Ti) show an inversely strong correlation. Occasionally, an element would behave slightly different between various kinds of statistical correlations (e.g. Zn weakly correlated with the heavy metals).

Based on visual and mineralogical observations of the sediment, it is apparent that the majority of the heavy metal contamination is in a detrital form. This is further supported by the lack of precipitation type mineral (e.g. goethite).

In conclusion, this study has shown that the outfall region and areas further downstream are highly contaminated with respect to heavy metals from the industrialized zone. Typically, fine-grained sediments are of great concern, because of their capacity to uptake and disperse contaminates, consequently moving them to areas of undesired impact. However, in this particular case it appears, the majority of the heavy metals are localized at the outfall region and slightly downstream. At some distance downstream, the heavy metals return to lower background levels. Accordingly, this suggests that the contamination is settling out of suspension and is redepositing in close proximity to the source. With this information, in terms of dredging the river, it is likely that resuspension of heavy metals will occur, however, may not prove to be as mobile as initially anticipated. It was observed that

a large amount oil existed in the river sediment, this study did not attempt explanation of organic influences on the sediment nor the surrounding environment.



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Appendix 1

Geochemical Data of Welland River Sediments



Appendix 1: Geochemical data of partial extraction analysis for dry sleved samples (All data in ppm).

ర	28.8	30.4	325.0	271.6	679.3	722.2	1140	6943	5589	7079	4251	0.86	2607	5136	9730	8010	6447	7893	4976	1109	1064	158.3	1076	1408	1988	1590	779.0	1507	763.1
П	194.8	101.1	539.7	147.4	213.7	113.0	239.4	217.2	264.1	222.2	229.0	235.2	147.9	202.4	117.6	204.5	320.3	227.5	153.9	170.9	163.2	313.2	185.8	183.1	165.9	152.9	193.3	176.2	184.2
Mn	350.5	245.7	359.5	191.4	292.4	255.3	366.6	3726	4529	3709	2976	466.2	3566	4028	3852	3662	3788	3887	3502	932.8	700.8	266.0	639.9	944.5	1041	729.9	518.8	579.1	580.2
Mg	6824	6229	12435	7896	9830	7500	11043	12577	6729	13313	10722	10922	9915	7160	4693	6711	4734	4976	8779	9710	9770	8039	9875	9194	10397	8745	8863	10244	9558
Иe	38722	30867	34331	31969	47094	34776	50913	271916	452071	289228	245572	31191	373519	376034	393908	394453	400992	421849	395458	102043	94116	31072	73159	108105	124843	116851	79547	94254	77325
Ba	141.2	111.2	102.7	125.1	128.8	143.4	129.5	381.6	385.0	353.9	171.4	89.44	278.7	426.2	311.2	1119	1219	2267	202.3	133.1	146.9	64.4	174.9	9.691	207.4	120.7	127.4	154.1	103.3
Al	17680	19766	12017	18556	16034	18966	16374	8832	9529	9705	10747	13311	8834	7439	5887	7190	8557	7814	9858	16504	19640	10373	16014	16540	15492	16189	14853	17962	15611
Sr	58.62	45.15	40.37	34.54	37.92	40.63	40.13	83.62	61.25	92.53	60.59	78.07	09.79	56.54	61.81	76.93	40.07	69.69	29.74	35.26	38.40	91.19	71.45	54.89	54.89	36.82	86.96	54.68	37.34
Zn	115.4	103.9	161.1	244.2	258.8	194.8	213.0	1435	549.5	1342	1122	110.1	858.5	613.5	497.9	345.3	412.6	384.8	255.8	232.5	229.9	113.4	284.0	290.6	234.1	211.6	549.1	233.4	200.8
Y	15.92	13.93	12.61	12.59	11.78	11.90	11.91	5.92	4.54	5.83	5.43	13.76	4.96	4.07	4.34	4.37	4.52	4.03	5.93	10.21	11.66	9.32	11.27	10.01	11.03	10.28	10.28	11.92	11.01
Λ	33.60	38.24	37.58	37.36	61.72	43.77	55.64	130.4	100.9	127.3	89.16	30.72	103.7	109.4	151.9	122.7	114.8	126.8	101.0	52.11	53.23	27.35	51.90	57.11	90.89	64.34	48.08	63.66	45.93
Sc	4.98	5.15	3.49	4.67	4.37	4.68	4.24	2.59	2.70	2.81	2.51	4.17	2.66	2.18	2.29	2.29	2.62	2.34	3.00	4.21	4.95	2.83	4.26	4.24	4.18	4.17	3.94	4.63	4.11
Z	30.17	33.53	168.0	161.7	383.3	211.8	421.9	5642	5039	9209	3646	109	6654	5159	7721	7389	4120	5825	2827	653.8	555.6	113.0	725.5	1329	1199	907.3	670.1	855.7	443.4
Cn	31.88	30.28	94.73	55.68	104.0	64.58	97.48	1198	948.6	1216	794.5	36.1	1175	810.2	851.3	477.5	705.4	569.1	443.9	132.4	127.4	37.77	132.1	140.6	171.1	162.1	110.8	140.4	94.43
ပိ	10.32	10.09	15.14	15.79	28.47	16.58	24.29	118.4	117.8	126.8	79.90	11.30	136.8	121.5	225.4	130.9	95.47	115.5	86.16	27.91	26.41	9.59	31.43	40.88	42.97	37.97	35.98	37.49	23.26
Be	0.92	1.03	0.59	0.78	0.70	0.82	69.0	0.35	0.33	0.40	0.39	0.63	0.33	0.27	0.27	0.26	0.30	0.26	0.34	0.64	0.77	0.53	0.82	0.81	0.82	0.64	0.62	0.72	0.63
locality		2	3	4	2	9	7	00	6	10	11	12	14	15	91	17	18	61	20	21	22	23	24	25	26	27	28	29	30
Sam	200QL	201QL	134QL	130QL	120QL	114QL	113QL	116QL	117QL	118QL	119QL	1040L	102QL	101QL	1050L	123QL	124QL	125QL	126QL	127QL	128QL	144QL	143QL	142QL	140QL	139QL	138QL	136QL	135QL

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Appendix 1: Geochemical data of partial extraction analysis for wet sleved samples (All data in ppm).

ර්	37.8	259.0	267.1	2380	137.8	4916	3791	3399	54.6	2983	4904	4443	3817	6046	7424	491.7	1855	115.8	1133	1844	596.2
E	126.4	125.9	338.7	75.1	113.0	257.5	163.6	199.0	218.1	144.0	202.5	148.0	126.1	435.8	248.0	103.4	186.2	339.4	163.7	188.0	190.5
Mn	233.8	195.5	362.1	3263	308.6	4086	2179	2855	479.7	2720	4057	3901	2867	3848	3084	426.0	632.5	259.7	607.7	1040	459.9
Mg	7350	8231	13246	9656	6686	6883	8531	11121	10780	11850	8485	7583	6989	5127	2488	10147	12032	8013	11399	9381	10188
Fe	33638	33380	38155	174744	39207	386371	156234	206357	33392	254820	355600	354541	276143	389843	332741	67184	84588	31337	73075	150495	65048
Ва	153.8	134.4	125.4	113.5	122.9	411.6	215.4	192.8	105.1	327.0	541.4	318.0	477.8	662.3	2311	170.5	397.6	84.7	257.1	134.5	135.2
Al	24314	19268	15274	13486	18937	10427	7039	12193	15506	12071	8645	8392	9222	8453	8176	19007	21545	13003	20539	14302	17820
Sr	46.96	35.53	47.43	41.10	50.32	64.70	58.22	67.79	87.17	77.18	69.28	64.75	66.39	68.99	75.55	29.60	61.22	52.88	68.51	37.60	43.56
Zn	121.3	270.4	209.5	344.5	209.5	0.089	972.2	1352	119.0	1218	915.6	9.469	1502	477.5	490.3	1041	386.7	106.4	410.3	195.5	260.2
Y	16.37	12.82	15.22	8.54	11.77	4.64	3.75	5.80	15.04	00.9	4.61	4.51	6.38	3.96	4.32	11.59	14.30	10.62	14.26	11.19	12.77
Λ	41.92	36.67	42.51	253.9	37.50	91.27	71.64	85.12	32.54	79.21	110.5	87.14	116.2	144.7	124.1	55.01	64.76	30.41	57.35	88.89	47.14
Sc	6.19	4.83	4.32	3.73	4.81	2.56	1.78	2.59	4.73	2.87	2.21	2.23	2.59	2.19	2.16	4.97	5.49	3.30	5.32	3.94	4.61
Z	39.18	158.4	147.9	2349	114.3	4846	3542	3361	72.21	6124	5458	4783	5362	4421	5149	593.1	889.2	86.46	6.689	1009	360.9
Cu	38.81	70.02	78.54	1234	88.18	7.786	877.7	948.2	35.63	1453	1002	978.6	1654	829.9	565.6	162.4	212.4	41.85	176.5	176.6	127.6
ပ္ပ	13.11	14.87	16.72	68.19	11.90	92.70	65.29	64.73	9.81	108.5	110.1	91.47	131.1	103.0	87.00	30.80	36.03	11.32	37.16	41.10	22.33
Be	1.12	0.92	0.82	1.23	0.89	09.0	0.46	0.65	08.0	0.65	0.61	0.54	0.72	99.0	0.61	86.0	0.97	09.0	0.95	0.64	0.79
ocality	1	3	4	9	7	00	6	10	11	12	13	14	15	16	17	19	21	24	26	27	28
Sam	201PL	130PL	134PL	112PL	121PL	117PL	118PL	119PL	104PL	102PL	101PL	108PL	105PL	122PL	125PL	129PL	145PL	144PL	143PL	139PL	135PL

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the detection limit of the instrument and the limit of quanitification of the method. The data in the following pages does not exclude the data below the HACL, These limits are summarized in the table below. Utilize accordingly.

Elemental	assessment	Elemental assessment of data by ICP-OES.	ICP-OES.			
Element	MDL	T00	Range Required	HACL	AcceptableData	Environments for which
	(mdd)	(mdd)	(mdd)	(mdd)	(2xHACL) (ppm)	Data is suitable.
Sc	1	3	1-27	2	>3	Background & Anthropogenic
Y	2	9	3-42	7	>14	Background only
Sr	1	3	13-340	52	>100	Background & Anthropogenic
Co	5	15	9-340	4	8<	Background & Anthropogenic
Cu	5	15	33-3900	21	>42	Background & Anthropogenic
Mo	9	18	4-1400	7	>14	Anthropogenic only
i.	2	9	27-12000	91	>32	Background & Anthropogenic
Ç	NA	NA	48-34000	21	>41	Background & Anthropogenic
Λ	5	15	48-400	24	>48	Background & Anthropogenic
W	35	105	0-600	8	>15	Anthropogenic only
Zn	2	9	68-3100	76	>150	Anthropogenic only
ΑΙ	NA	NA	3577-87430	11752	>23500	Anthropogenic only
Ва	NA	NA	89-5300	82	>160	Background & Anthropogenic
Fe	NA	NA	11000-260000	9200	>18400	Anthropogenic only
Mg	NA	NA	2288-44300	2076	>4150	Background & Anthropogenic
Mn	NA	NA	255-7800	95	>190	Background & Anthropogenic
Ti	NA	NA	192-11000	286	>1970	Anthropogenic only

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Appendix 1: Geochemical deta of total extraction analysis for dry sleved samples (All data in ppm).

Sample	L0c	Be	ပိ	J	Ż	Sc	>	<b>X</b>	Zn	Sr	A	Ba	Fe	Mg	Mn	Ë	Ċ
2000	-	1.98	16.83	36.5	40.0	12.92	102.90	23.08	151.7	196.80	78962	535.8	41247	12204	445.8	4701.8	69.2
2010	2	2.16	17.46	33.6	44.3	14.04	117.70	19.74	140.8	172.90	86510	522.4	43953	11899	316.0	5518.2	17.1
1340	~	1.55	21.76	127.4	221.6	11.57	106.89	25.63	237.2	221.01	70857	551.9	53395	22829	878.6	5702.0	513.4
1300	4	1.96	21.46	71.8	187.8	12.16	105.60	17.98	303.5	165.00	78757	522.2	43987	13291	268.3	4618.5	397.7
1200	\$	1.86	39.75	136.9	530.3	11.92	144.30	19.75	335.7	161.00	73296	501.0	65100	91091	427.8	4713.3	1240.7
1150	9	1.89	23.68	61.2	204.3	12.81	114.56	21.57	209.8	16.902	81610	596.2	52400	18282	710.0	5655.9	549.7
1140	7	1.82	21.30	72.6	248.7	11.81	109.90	16.42	223.6	143,30	73982	467.7	43739	12026	314.9	4193.4	992.2
1130	00	1.73	30.48	112.4	452.6	10.98	122.40	17.08	254.7	158.60	70161	438.5	60820	16324	461.9	4137.0	1641.5
1120	6	0.98	110.21	1127.6	2898.0	4.58	313.23	8.59	277.5	66.62	22263	209.5	258114	9273	10438.6	1351.5	7825.5
1110	10	0.71	105.79	1043.8	2300.3	6.44	239.47	14.19	746.4	157.62	33348	405.2	275464	16728	12897.8	2022.4	10240.7
1210	Ξ	2.00	22.15	84.6	197.5	13.58	131.47	09.61	239.6	213.39	85080	9.078	08609	19289	748.5	5111.1	458.5
1160	12	98'0	150.60	1433.0	6907.0	5.48	205.70	10.24	1755.0	137.40	29960	567.8	281664	16503	4843.9	1750.7	15599.4
1170	13	0.59	141.60	971.5	5820.0	3.51	141.60	5.84	594.8	80.13	18961	455.0	316757	8051	4436.5	1024.0	15162.0
1180	14	1.56	260.91	2403.0	12204.0	8.87	336.40	17.43	2542.9	243.93	44473	877.5	453438	30153	10284.7	2863.7	24539.4
1190	15	0.85	103.20	952.1	4434.0	4.86	143.30	8.74	1354.0	101.00	30387	304.4	243198	14225	3716.7	1519.7	9369.6
1040	16	1.57	15.57	49.7	117.9	6.67	80.34	19.80	134.3	189.10	64340	422.2	39304	14879	520.6	4498.9	228.4
1020	17	0.72	173.60	1260.0	8074.0	4.51	158.70	7.19	922.1	96.84	23241	373.8	304564	11881	3664.5	1172.2	15049.4
1010	18	0.59	155.80	900.4	6228.0	3.17	172.00	5.71	684.6	81.54	19343	524.5	272933	9396	4360.5	896.5	16765.8
1080	19	0.49	149.00	941.4	6229.0	2.89	144.10	4.78	494.1	64.94	16230	307.8	348772	7482	4651.6	730.5	16607.0
1050	20	1.12	299.49	1013.2	10266.5	4.34	249.10	8.56	626.0	109.38	18985	475.3	323804	7581	6503.1	1037.2	24357.8
1220	21	0.38	133.30	634.7	3929.0	2.50	154.70	3.58	238.9	49.42	12775	429.5	294531	4253	4345.9	822.7	14123.8
123Q	22	0.59	151.70	484.2	8022.0	3,46	177.40	5.63	340.7	106.00	20732	8254.9	235940	8928	3419.5	902.0	22765.1
1240	23	0.70	133.60	6.108	5456.0	4.52	180.70	6.59	479.2	72.25	24502	1501.5	267489	6836	3968.3	1539.6	17459.1
1250	24	0.57	140.20	609.7	6562.0	3.53	182.70	5.50	404.7	27.06	22816	2577.1	254294	7507	3744.4	956.7	20129.0
126Q	25	0.83	106.90	476.0	3473.0	5.38	149.50	8.32	275.9	71.36	30494	338.9	224306	8100	6418.1	1747.7	13118.9
1270	26	1.82	42.81	1.72.1	912.5	11.77	145.07	18.12	325.2	170.13	73232	555.4	129267	17552	1868.8	4646.7	3823.9
128Q	27	1.80	34.44	150.8	0.899	11.72	122.90	16.07	275.4	136.20	72193	465.6	88106	14911	809.7	4159.9	2418.4
1440	28	1.33	16.12	43.3	125.7	7.98	70.79	14.80	138.8	184.80	92695	406.1	39483	11645	335.4	3941.8	296.8
1430	29	1.72	46.49	162.0	952.6	10.97	125.90	17.08	351.3	173.00	66427	492.8	91294	15382	826.6	3874.9	2823.8
1420	30	1.81	63.96	170.9	1729.0	11.70	139.60	17.23	373.5	156.50	26689	500.7	136557	15073	1187.9	3826.6	3885.6
1400	31	1.67	60.79	206.8	1530.0	10.81	154.40	16.64	283.6	152.00	63860	527.6	132393	15648	1238.8	3280.5	6211.8
1390	32	1.70	55.12	209.3	1243.0	10.01	147.80	16.42	269.9	139.50	65559	442.4	125857	14241	902.5	3446.7	5280.5
138Q	33	19.1	49.19	144.4	895.6	10.22	117.10	16.77	664.9	208.00	62786	118.0	98603	14084	0.189	3605.6	2214.5
136Q	34	1.78	52.68	171.2	1110.0	11.32	144.00	17.39	280.3	157.70	68043	479.9	115355	15339	737.7	3828.7	4696.1
135Q	35	1,61	32.35	124.0	598.0	10.39	110.00	16.88	279.9	151.00	65454	430.9	80535	14589	711.1	3886.6	2250.3
117QD		19.0	142.70	999.5	5802.0	3.63	140.30	60.9	611.7	81.79	20265	465.6	302480	8274	4914.7	1026.7	14689.7
120QD		1.81	38.60	130.5	500.7	11.42	139.60	19.10	322.7	154.70	69304	483.3	62546	15480	413.4	4531.0	1311.6
123QD		0.59	157.60	509.1	8561.0	3.56	179.40	5.78	346.1	105.80	20884	8713.2	235728	9011	3452.1	931.6	24030.3
124QD		89.0	129.90	785.0	5264.0	4.41	174.90	6.47	471.4	71.50	24070	1486.7	253093	6748	3823.2	1506.4	16517.9
138QD		1.58	48.49	141.0	872.1	10.07	115.00	16.80	650.1	205.50	61647	182.6	08086	13771	6.699	3563.6	2124.2
139QD		1.71	55.45	216.8	1220.0	11.08	149.50	16.45	278.7	140.30	66012	445.7	128521	14408	917.0	3325.0	5353.2
142QD		1.75	61.21	165.3	1595.0	11.38	134.00	16.58	361.0	152.10	89029	489.4	127149	14718	1125.4	3689.3	2779.6
201QD		2.19	17.64	34.0	44.2	14.00	118.40	20.14	141.8	172.60	86458	521.9	43813	11854	317.1	5289.2	77.0

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Appendix 1: Geochemical data of total extraction analysis for wet sleved samples (All data in ppm).

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ڻ	95.	325.	361.	8225.	394.	15818.	14924.	9446.	139.	18225.	16430.	19454.	15145.	21347.	949.	4338.	181.	2635.	5845.	1922.	16565.	3874
Ξ	6928.1	4863.3	5582.0	3792.2	5590.0	1884.2	2501.6	2202.3	4781.6	1183.5	1369.8	2083.7	1127.4	1599.6	5417.3	5675.7	5454.0	5653.7	4182.8	2400.0	1742.1	5851.4
Mn	495.5	262.0	765.7	6658.5	2999	7332.9	7033.8	5423.0	510.0	5456.8	7386.3	8.0619	4456.1	5756.5	852.9	1294.5	511.6	1212.1	1937.6	951.1	7363.6	1337.6
Mg	15127	13724	21028	18502	18478	12563	22705	17751	15008	11337	11499	12525	5998	9358	18641	21608	12547	20002	15628	18200	14325	22442
Fe	48937	43245	52065	22739	58400	349136	02820	49898	41064	102532	142123	16321	13897	328546	91653	113001	40134	98319	163548	84402	336712	116321
				431.2						, ,		. ,										
				63760																		
Sr	184.31	155.50	196.47	130.68	185.52	122.87	177.44	127.13	191.30	95.20	99.22	128.69	73.78	114.64	175.06	178.32	193.59	191.32	155.43	188.60	126.74	184.50
				502.9																		
				17.52																		
>	141.87	107.10	112.43	537.88	133.38	195.60	235.01	175.56	90.21	189.40	185.83	266.54	199.40	234.57	166.65	179.32	81.30	163.84	160.52	135.60	193.18	183.47
Sc	17.01	12.64	12.42	11.95	14.69	5.94	7.78	6.52	11.38	3.89	4.75	7.05	3.20	5.30	15.31	15.71	9.43	15.00	10.63	13.18	5,83	16.07
Z	9.99	174.7	176.3	3757.9	169.0	6584.5	7998.5	4827.8	80.9	7044.0	6851.9	8285.2	5108.0	6922.6	773.1	1205.0	93.9	930.9	1339.2	475.1	6985.0	1227.5
Ö	41.8	6.18	121.3	1727.3	112.2	1320.6	2012.5	1215.9	40.9	1101.0	1243.3	2193.8	827.5	671.9	200.6	261.9	39.7	241.6	210.3	142.0	1389.6	256.2
ပိ	19.47	2163	21.03	126.58	21.65	155.46	170.48	110.11	15.69	177.20	160 36	230.92	160 10	144 44	45.52	48.71	14.99	49.68	56.52	30.32	161.46	50.67
Be	2.32	1 99	1.74	3.02	2.07	0.83	1.26	1.03	1.75	0.68	0.65	1 44	0.58	1.06	2.34	2.44	1.24	2.28	1.68	1.95	0.85	2.53
Loc	-	ح) .	. 4	. 9	7	00	0	10	=		14	15	2 9	17	10	21	24	96	27	28		
Samule	2019	1300	134P	112p	121P	117P	118P	1199	104P	1010	1080	105p	122P	125P	129P	145p	144P	143P	1395	1355	CIGROT	145PD

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Appendix 1: Geochemical data of grain size separated samples (All data in ppm)

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Appendix 2

Microwave Digestion Technique



## Appendix 2:

# Microwave Digestion Technique

Experiments with microwave digestion techniques were performed at Ministry of the Environment under supervision of Dr. B. Campbell. Several parameters were experimented with different reagents, times and pressures of the system.

#### Calibration:

To calibrate the working power of the microwave system, a volume of water (1L) was heated in the microwave system for 1 minute. The measurement was repeated 3 times. An initial and final temperature was obtained. Following equation 1, the working power of the microwave system was calculated.

(1) Working Power (W) = 70 X (Tf - Ti), where Tf - final temperature, Ti - initial temperature.

Table 1.

	Ti(°C)	Tf(°C)	W(watts)
trial 1	19	37	1270
trial 2	19	36	1190
trial 3	19	36	1190
expected power			1200

The digestion was performed at half the power of this reading (600 W).

Sample Digestion Procedure: (Table 2,3a,4a).

- -sample weighed in Teflon bomb (0.2-0.3g)
- -HNO<sub>3</sub> or aqua regia added (varying amounts)
- -inserted into microwave for digestion
- -once completed the bombs were removed and cooled in water
- -pressure released and bombs uncapped
- -solution transferred and deionized water added to bring up to 50 ml
- -residue was filtered and retained
- -colour of solution and residue recorded

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Further testing of the residues and solutions included:

- 1. XRD of the residue to identify mineral present
- 2. chemical analyses of the solutions to identify % recovery of standard reference materials

#### Results of these tests:

- 1. XRD identified only minerals with a silicate structure. There were no oxides or carbonates present
- 2. chemical analyses indicated there was a recovery of approximately 70% (Table 3a,3b,4a,4b).

### Pros:

-all minerals except silicate minerals went into solution

### Cons:

- -Ministry of the Environment would not authorize the use of HF in bombs. The certified values for standard reference materials that were used were results for total extraction, not partial extraction. Without the use of HF in the digestion, the analyses could not be validated for accuracy.
- -The final solutions contained an acid concentration of approximately 30%. This was not optimal as a final working solution for ICP analysis. Dilution of the solutions made some elements undetectable.
- -Time allocation needed for method development, validation and generation of chemical results was limited.

Table 2. Experiments with Microwave Digestion, Run 1 through 3: Using program 4 for steps 1 through 4, program 5 for step 5.

Time

2

	3				
2	**	200	2		
3	11	400	2		
4	n	600	5		
5	**	600	4		
vessel	sample	sample weight	reagent volume	colour solution	colour residue
runl					
1	101a	.3088	8ml	green	black grains
2	101b	.2998	19	green	**
3	101c	.3019	r	green	brown pwd
4	101d	.3007	FF .	green	11
5	101e	.3024	11	yellow-green	н
6	108a	.3086	Ĥ	yellow-green	black grains
Run2					
1	127c	.3022	8ml	lg yellow	lg brown
2	11	.3018	**	n	ii
3	н	.3034	19	н	11
4	п	.3011	19	clear	11
5	**	.2990	11	tt.	Ħ
6	н	.3018	11	lg yellow	11
run3					
1	127c	.2042	8ml	clear	lg brown
2	11	.1993	FF	39	99
3	11	.2026	11	19	"
4	11	.1998		11	"
5	H	.1992	11	19	"
6	11	.2015	Pt .	h	11

note: no chemical analyses for these samples.

Step

1

Reagent

HNO<sub>3</sub>

Power



Table 3a.

Run 1 and 2: Using program 4 for steps 1 through 4, program 5 for step 5.

Step	Reagent	Power	Time
1	HNO <sub>3</sub>	120	2
2	11	200	2
3	11	400	2
4	# -	600	5
5	н	600	4

vessel	sample	sample weight	volume	solution colour	residue colour
runl					
1	blka	0	8ml	clear	lg brown
2	blkb	0	11	н	н
3	2704	.1984	89	12	11
4	2704d	.1966	19	19	н
5	1646	.2009	11	19	11
6	1646d	.2004	11	11	н
run2					
1	blk 1	0	8ml	clear	lg brown
2	blk2	0	99	99	99
3	blk3	0	19	99	H
4	blk4	0	н	-11	n
5	blk5	0	н	н	11
6	blk6	0	"	11	H

remarks: 0.200g in 50ml = 250 Dilution Factor.



Table 3b.

Results of initial microwave digestion of blanks, NIST-2704 and NIST-1646. Digestion reagent was HNO<sub>3</sub> (achieved approx. 80-90% digestion).

Smps	Be	Co	Cu	Ni	V	Y	Zn	Mo	Sc
blka	0	2.0	2.37	9.84	0.87	0.49	16.3	3.6	1.18
blkb	0	2.29	2.10	9.13	0.99	0.57	17.3	3.43	1.35
2704	0.92	12.78	95.16	48.32	37.23	16.23	449.5	0.88	5.8
2704D	0.86	13.05	85.14	47.84	35.19	15.76	463.3	1.73	6.1
1646	0.91	10.06	17.85	34.87	52.89	15.35	121.7	0.04	7.5
1646D	0.88	10.03	18.15	34.56	49.65	14.94	117.9	0.12	7.1
blk1	0	2.8	2.5	10.48	1.61	0.77	1.78	4.14	1.55
blk2	0	1.83	0.423	9.36	0.13	0.28	1.26	3.5	1.1
blk3	0	1.97	1.1	10.18	0.44	0.48	1.83	3.37	1.13
blk4	0	1.95	0.91	10.15	0.32	0.36	1.39	2.96	1.03
blk5	0	1.88	1.17	10.07	0.30	0.26	1.94	3.44	1.18
blk6	0	2.21	0.72	10.04	0.28	0.31	1.67	4.26	1.12
Certified value 2704	NA	14.0	98.6	44.1	95.0	NA	438.0	NA	NA
accuracy 2704 (%)	NA	8.7 6.8	3.5 13.6	9.6 8.5	60.8 62.9	NA	2.6 5.8	NA	NA
Certified value	1.5	10.5	18.0	32.0	94.0	NA	138.0	NA	10.8
accuracy 1646 (%)	39.3 41.3	4.2 4.5	0.8	9.0 8.0	43.7 47.2	NA	11.8	NA	30.5

Accuracy (%) =  $(known \ value - measured \ value)$  X 100 (known value)



Table 4a. Experiments with Microwave Digestion. Run 1 through 3: Using program 4 for steps 1 through 4.

vesse	smp	smp wt	reagent	Vol	soln	residue	step	Pwr	time
1	blk1	0	HNO <sub>3</sub>	8ml	clear	lgbrown	1	120	2
2	blk2	0	- 19	8ml	11	н	2	200	2
3	2704	0.1504	11	8ml	11	н	3	400	2
4	2704D	0.1520	19	8ml	**	11	4	600	9
5	1646	0.1502	n	8ml	**	11			
6	1646D	0.1477	**	8ml	"	"			
1	blk1	0	HNO <sub>3</sub>	8ml	11	"	1	120	2
2	blk2	0	11	11	11	"	2	200	2
3	2704	0.1503	п	**		"	3	400	2
4	2704D	0.1499	н	**	19	"	4	600	9
5	1646	0.1499	H <sub>2</sub> O <sub>2</sub>	1ml	**	"	5	160	3
6	1646D	0.1509			"	11			1
1	blk1	0	HNO <sub>3</sub> +HCl(3:1)	8ml			1	120	2
2	blk2	0	99	н			2	200	2
3	2704	0.1509	17	н			3	400	2
4	2407D	0.1489	н	**			4	600	9
5	1646	0.1525							
6	1646D	0.1514							

remarks: .150g in 25ml = 166.67 Dilution Factor.

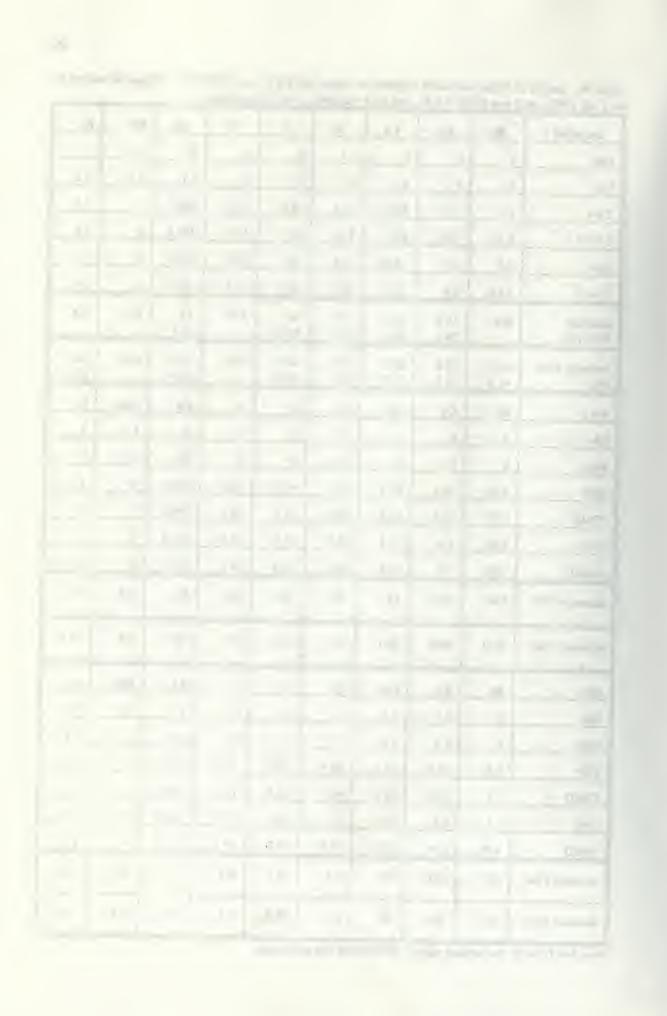


Table 4b. Results of initial microwave digestion of blanks, NIST-2704 and NIST-1646. Digestion reagent for

set 1 was HNO<sub>3</sub>, set 2 was HNO<sub>3</sub> + H<sub>2</sub>O<sub>2</sub>, and set 3 was HNO<sub>3</sub> + HCl (aqua regia).

Smps/Set 1	Be	Co	Cu	Ni	V	Y	Zn	Mo	Sc
blka	0	0	0	2.2	0	0	2.1	1.2	0
blkb	0	0	0	2.0	0	0	2.1	4.0	0.2
2704	0.9	10.3	82.6	39.4	43.6	16.5	403.8	0	5.4
2704D	0.95	10.0	80.8	37.9	45.4	16.0	383.6	0	5.8
1646	0.8	6.9	12.75	24.6	53.7	13.9	107.3	0	6.0
1646D	0.88	6.8	12.6	25.5	56.4	14.5	109.0	0	6.4
accuracy 2704 (%)	NA	26.4 28.6	16.2 18.1	10.7 14.1	54.1 52.2	NA	7.8 12.4	NA	NA
accuracy 1646 (%)	46.7 41.3	34.3 35.2	29.2 30.0	23.1 20.3	42.9 40.0	NA	22.2 21.0	NA	44.4 40.7
Set 2	Be	Со	Cu	Ni	v	Y	Zn	Mo	Sc
blka	0	0	0	3.5	0	0	0	1.5	0
blkb	0	0	0	2.6	0	0	3.6	1.4	0
2704	0.95	10.4	83.5	39.3	44.6	16.6	394.5	0	5.8
2704D	0.97	10.8	83.5	40.6	43.6	16.5	399.5	0	5.9
1646	0.93	7.4	13.3	28.8	58.7	15.5	115.8	0	7.1
1646D	0.88	7.3	12.8	28.0	56.6	14.9	150.4	0	6.7
accuracy 2704	NA	25.7	15.3	10.9	53.1	NA	9.9	NA	NA
accuracy 1646	38.0	29.5	26.1	10.0	37.6	NA	16.1	NA	34.3
Set 3	Be	Co	Cu	Ni	v	Y	Zn	Mo	Sc
blka	0	0.5	1.5	2.7	0	0	3.1	1.4	0.4
blkb	0	0.4	1.5	3.4	0	0	2.9	1.3	0.4
2704	1.43	11.9	91.4	44.9	70.2	18.7	432.8	0	9.6
2704D	1.5	12.1	88.6	44.3	72.5	19.5	425.0	0	10.
1646	1.1	8.8	17.1	31.7	66.0	15.7	123.8	0	8.4
1646D	1.05	8.9	17.0	32.5	65.3	15.7	122.2	0	8.5
accuracy 2704	NA	15.0	7.3	1.8	26.1	NA	1.2	NA	NA
accuracy 1646	26.7	16.2	50.	0.9	29.8	NA	10.3	NA	22.2

Note: See Table 3b. for certified values. All chemical data are in ppm.



## Appendix 3

pH Data of the Water of the Welland River



Appendix 3: pH of water of the Welland River.

Sample Site	pH Reading
101	7.34
102	7.85
103	8.10
104	7.80
105	7.58
108	8.13
Average	7.80

<sup>\*</sup>Temperature of water at time of measurement 20°C.

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Appendix 4

Sample Key



## Appendix 4: Sample Key

## Particles sizes (µm)

A - 63-45

B - 44-32

C - 31-23

D - 22-13

E - 12-11

F - <11

P - bulk of grain sizes  ${<}63~\mu m$  -wet sieving

Q - bulk of grain sizes <63 µm -dry sieving

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Appendix 5

Field Sample Description Data



## Appendix 5: Fleld Sample Description Data of Welland River Samples

Sample ID	Location	Description	pН
101	at outfall	lots of oil, large pebbles	7.34
102	at outfall	lots of oil, varying grain size	7.85
103	at outfall	lots of oil, foul smell, very black	8.10
104	at outfall	medium brown colour, fine-grained sediment	7.80
105	at outfall	lots of oil,	7.58
106	at outfall	rock sample	
107	at outfall	rock sample	
108	at outfall	lots of oil, metal, black sample	8.13
110	50m upstream	fine brown sediment, oil present	
111	100m upstream	fine brown sediment, oil present	
112	125m upstream	fine brown sediment, oil present	
113	175m upstream	fine brown sediment, oil present	
114	150m upstream	fine brown sediment, oil present	
115	200m upstream	fine brown sediment, oil present	
116	outside boom at outfall	lots of oil in water and sample	
117	at second reef	metal and oil in sample	
118	slightly inside boom at outfall	lots of oil in water and sample	
119	outside boom at outfall	lots of oil in water and sample	
120	225m upstream	strong organic smell	
121	10m upstream	very organic sample, alot of algae	
122	30m downstream	black mud with very strong smell	
123	50m downstream	strong organic smell, lots of oil in sample	
124	70m downstream	strong organic smell, lots of oil in sample	
125	100m downstream	strong organic smell, lots of oil in sample	
126	200m downstream	strong smell	
127	400m downstream	strong smell	
128	at sewage disposal plant	strong smell	
129	100m downstream sewage	strong smell	
130	20m upstream from outfall 2	fine brown sediment	
131	50m upstream from outfall 2	fine brown sediment	
132	70m upstream from outfall 2	fine brown sediment, metal, zebra mussels in sample	
133	150m upstream from outfall 2	fine brown sediment, metal chunks, glass in sample	
134	200m upstream from outfall 2	fine brown sedment	
135	400m down from Dump	mud sample	
136	300m down from Dump	mud sample, oil and metal present	
137	275m down from Dump	mud sample, oil and metal present	
138	225m down from Dump	mud sample, oil and metal present	
139	170m down from Dump	mud sample, oil and metal present	
140	50m down from Dump	mud sample, metal and alot of oil present	
141	under Atlas Steel Dump	mud sample, metal and alot of oil present	
142	120m up from Dump	mud sample, brown in colour	
143	200m up from Dump	mud sample, brown in colour	
144	400m up from Dump	mud sample, brown in colour	
145	500m up from Dump	mud sample, brown in colour	









